

US EPA ARCHIVE DOCUMENT

**PROPOSED  
TOTAL MAXIMUM DAILY LOAD (TMDL)**

**For**

**Dissolved Oxygen and Nutrients**

**In**

**McKay Creek  
(WBID 1633B)**

**November 2012**



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**SUMMARY SHEET for WBID 1633B****Total Maximum Daily Load (TMDL)****2009 303(d) Listed Waterbodies for TMDLs addressed in this report:**

WBID	Segment Name	Class and Waterbody Type	Major River Basin	HUC	County	State
1633B	McKay Creek	Class III Freshwater	Tampa Bay	3100207	Pinellas	Florida

**TMDL Endpoints/Targets:**

Dissolved Oxygen &amp; Nutrients

**TMDL Technical Approach:**

The TMDL allocations were determined by analyzing the effects of TN and TP concentrations and loadings on DO concentrations in the waterbody. A watershed model and estuary model were used to predict delivery of pollutant loads to the waterbody and to evaluate the in-stream impacts of the pollutant loads.

**TMDL Waste Load and Load Allocation**

Constituent	Current Condition		TMDL Condition		Percent reduction		
	WLA (kg/yr)	LA (kg/yr)	WLA (kg/yr)	LA (kg/yr)	WLA	LA	MS4
Total Nitrogen	--	6,838	--	4,445	--	35%	35%
Total Phosphorus	--	896	--	582	--	35%	35%

**Endangered Species Present (Yes or Blank):**

USEPA Lead TMDL (USEPA or Blank): USEPA

TMDL Considers Point Source, Non-point Source, or Both: Non-point

**Major NPDES Discharges to surface waters addressed in USEPA TMDL:**

Permit ID	Permittee(s)	County	Permit Type
FLS000005	Cities of Bellair Bluffs / Clearwater / Largo	Pinellas	Phase I MS4
FLS000007	City of St. Petersburg	Pinellas	Phase I MS4

## 1.0 INTRODUCTION

Section 303(d) of the Clean Water Act requires each state to list those waters within its boundaries for which technology based effluent limitations are not stringent enough to protect any water quality standard applicable to such waters. Listed waters are prioritized with respect to designated use classifications and the severity of pollution. In accordance with this prioritization, states are required to develop Total Maximum Daily Loads (TMDLs) for those water bodies that are not meeting water quality standards. The TMDL process establishes the allowable loadings of pollutants or other quantifiable parameters for a waterbody based on the relationship between pollution sources and in-stream water quality conditions, so that states can establish water quality based controls to reduce pollution from both point and nonpoint sources and restore and maintain the quality of their water resources (USEPA 1991).

The Florida Department of Environmental Protection (FDEP) developed a statewide, watershed-based approach to water resource management. Under the watershed management approach, water resources are managed on the basis of natural boundaries, such as river basins, rather than political boundaries. The watershed management approach is the framework FDEP uses for implementing TMDLs. The state's 52 basins are divided into five groups and water quality is assessed in each group on a rotating five-year cycle. FDEP also established five water management districts (WMD) responsible for managing ground and surface water supplies in the counties encompassing the districts.

For the purpose of planning and management, the WMDs divided the district into planning units defined as either an individual primary tributary basin or a group of adjacent primary tributary basins with similar characteristics. These planning units contain smaller, hydrological based units called drainage basins, which are further divided by FDEP into "water segments". A water segment usually contains only one unique waterbody type (stream, lake, canal, etc.) and is about 5 square miles. Unique numbers or waterbody identification (WBID) numbers are assigned to each water segment. This TMDL addresses WBID 1633B, which is a Group 5 waterbody located in the Anclote River/Coastal Pinellas County Planning Unit and is managed by the Southwest Florida Water Management District (SWFWMD). WBID 1633B is impaired for dissolved oxygen (DO) and nutrients.

## 2.0 PROBLEM DEFINITION

To determine the status of surface water quality in Florida, three categories of data – chemistry data, biological data, and fish consumption advisories – were evaluated to determine potential impairments. The level of impairment is defined in the Identification of Impaired Waters Rule (IWR), Section 62-303 of the Florida Administrative Code (FAC). The IWR is FDEP's methodology for determining whether waters should be included on the state's planning list and verified list. Potential impairments are determined by assessing whether a waterbody meets the criteria for inclusion on the planning list. Once a waterbody is on the planning list, additional data and information will be collected and examined to determine if the water should be included on the verified list.

The TMDL addressed in this document is being established pursuant to commitments made by the United States Environmental Protection Agency (USEPA) in the 1998 Consent Decree in the Florida TMDL lawsuit (Florida Wildlife Federation, et al. v. Carol Browner, et al., Civil Action No. 4: 98CV356-WS, 1998). That Consent Decree established a schedule for TMDL development for waters listed on Florida's USEPA approved 1998 section 303(d) list. The 2009 section 303(d) list identified numerous WBIDs in the Anclote River/Coastal Pinellas County Basin as not meeting WQS. After assessing all readily available water quality data, USEPA is responsible for developing a TMDL for WBID 1633B, depicted in Figure 2.1.

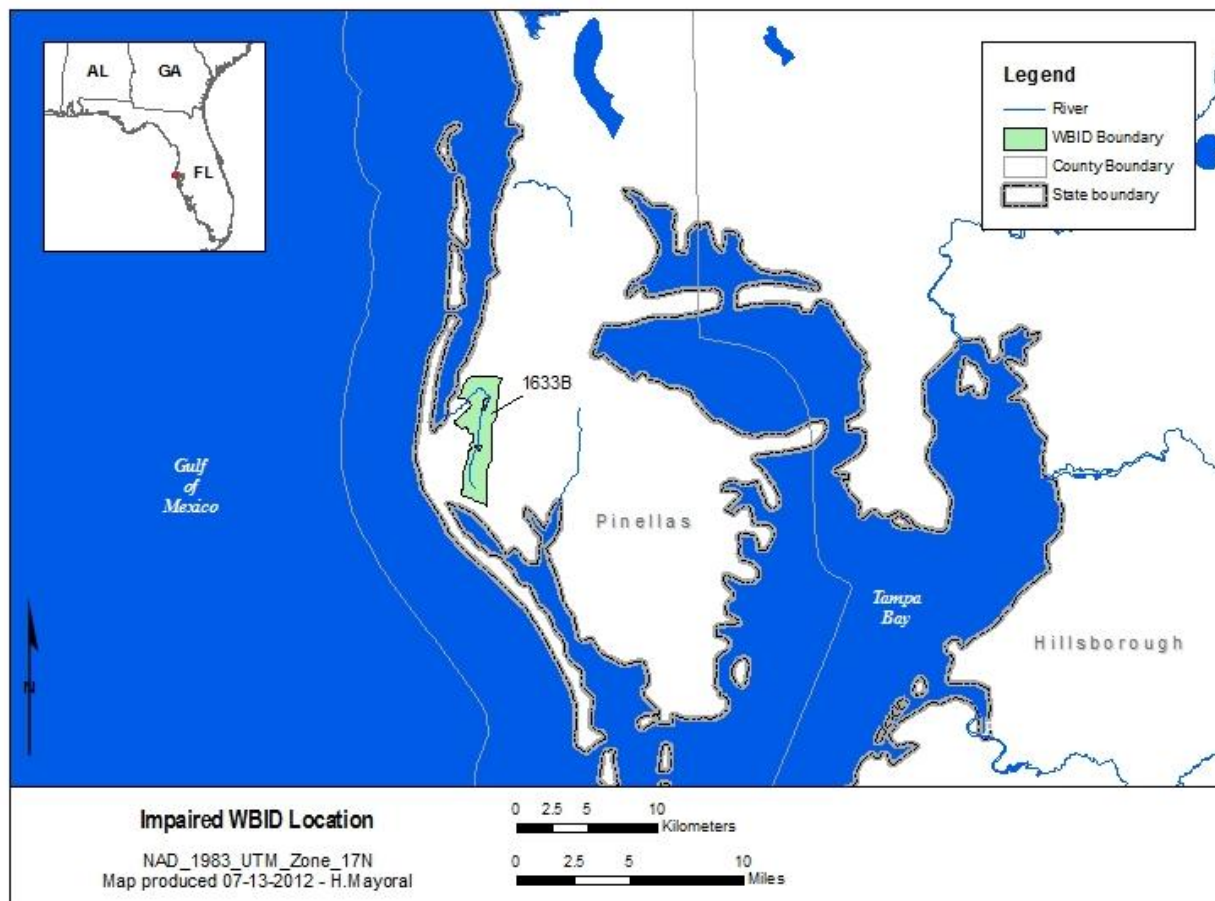


Figure 2.1 Location of the impaired WBID in the McKay Creek watershed

### 3.0 WATERSHED DESCRIPTION

The Tampa Bay Basin is located along the west coast of Florida, beginning just south of the Withlacoochee River in Citrus County and extending to Gulfport, Florida in Pinellas County, although it does not include Tampa Bay. Six major rivers are located within the watershed: the Crystal River, Homosassa River, Chassahowitza River, Weeki Wachee, the Anclote River, and the Pithlachascotee River (FDEP). The Brooksville Ridge marks the eastern boundary of the Springs Coast Basin, created by sands historically deposited during higher sea-levels, and which define the karst geology that is characteristic of the area (FDEP 2008). Three physiographic

regions with varying geology and topography are located within the Springs Coast Basin, the Coastal Swamps, the Gulf Coastal Lowlands, and the Brooksville Ridge. There are 194,500 acres in the basin dedicated to conservation, with approximately 141,350 acres, or 73 percent, sandwiched between the Gulf of Mexico and U.S. Highway 19. Conservation lands in the basin include 130,250 acres of state-owned lands, 18,500 acres of Southwest Florida Water Management District (SWFWMD)-owned lands, 3,500 acres of county-owned lands, and nearly 1,000 acres of privately owned lands (FDEP 2008).

Two main aquifers are found in Pinellas County, the surficial aquifer and the Floridan aquifer. The surficial aquifer system consists of undifferentiated sands, shell material, silts and clayey sands of varying thickness (Causseaux, 1985). The principal uses for the surficial aquifer in Pinellas County are irrigation, limited domestic use, and dewatering projects for mining and infrastructure installation (SWFWMD, 2006). The Floridan aquifer system consists primarily of highly permeable carbonate rocks and is separated into two principal zones consisting of the fresh potable water of the Upper Floridan aquifer and the highly mineralized water of the Lower Floridan aquifer (Causseaux, 1985). In Pinellas County, the Upper Floridan aquifer is the principal source of water and is used for industrial, mining, public supply, domestic use, and irrigation purposes, as well as brackish water desalination in coastal communities (SWFWMD, 2006).

An important feature of the area is karst topography. Watersheds located in karst regions are extremely vulnerable to contamination. Many of these karst features infiltrate the water table forming a direct connection between land surface and the underlying aquifer systems, allowing interaction between surface and ground waters (SWFWMD, 2002) increasing the threat of ground water contamination from surface water pollutants (Trommer, 1987). Potential sources of contamination include saltwater encroachment and infiltration of contaminants carried in surface water, direct infiltration of contaminants (chemicals or pesticides applied to or spilled on the land, fertilizer carried in surface runoff), landfills, septic tanks, sewage-plant treatment ponds, and wells used to dispose of stormwater runoff or industrial waste (Miller, 1990).

### **3.1 Climate**

The Tampa Bay Basin is located on the west coast of Central Florida and experiences a subtropical climate with hot, humid summers and mild, short winters. Average high temperatures in the summer are in the low-90s (°F), and average low temperatures in the winter are in the upper-40s (°F). An average of 52 inches of rain every year is received in this part of Central Florida, of which a greater percentage falls during the wet season (June through September) (SERCC 2012).

### **3.2 Hydrologic characteristics**

The McKay Creek watershed is located in the west central area of Pinellas County and includes part of the cities of Largo and Seminole and the Town of Belleair Bluffs. These areas are almost completely developed. Major land use types include low-, medium- and high- density residential, institutional, commercial lands, and recreational/open space (undeveloped and developed parkland). The watershed is within the Gulf Coastal Lowlands physiographic region, where soils are poorly drained and the water table is near land surface. Soils in this region are variable; they

range from excessively drained sands to moderate or poorly drained soils with a sandy subsoil (USDA, 2006). As a result of extensive changes of the land surface for development, large portions of this area have soils types characterized as Urban Land (SWFWMD, 2002).

McKay Creek (WBID 1633B) is a stream that is approximately six miles in length and channelized along most of its reach. The stream originates just south of 86<sup>th</sup> Avenue North in Pinellas County. The stream flows northerly, and then turns toward the west and then the southwest, where it drains to McKay Creek Tidal (WBID 1633), and then empties into Clearwater Harbor. Clearwater Harbor is separated from the Gulf of Mexico by barrier islands, and is hydrologically connected to the Gulf through occasional inlets. There are two lakes (with unique WBID numbers) located within WBID 1633B, Taylor Lake and the Walsingham Reservoir. Both are located along the main channel of the McKay Creek. Additional information about the hydrology of this area is available in the General Hydrology of the Middle Gulf Area, Florida (Report of Investigation No. 56), by the US Geological Survey (Cherry et al., 1970).

### **3.3 Land Use**

A majority of the land use in WBID 1633B McKay Creek is classified as developed land use (Figure 3.1 and Table 3.1). A total of 88 percent of the land use is classified as developed, and 73 percent of the total land use is classified as high intensity development. Combined forest land uses account for 5 percent of the total land use, with most forested land uses located along the riparian corridor of McKay Creek. Small areas of forested and non-forested wetlands are located in the headwaters, accounting for 2 percent of the total land use in the WBID. Agricultural land accounts for one percent of the total land use and is also located in the headwaters of McKay Creek. Open water accounts for 4 percent of the total land use.

The actual drainage area for McKay Creek varies from the WBID boundary (Figure 3.2 and Table 3.2). The United States Geological Survey National Hydrography Dataset was used to delineate the drainage area. Acreage and land use did not vary considerably between the drainage area and the WBID boundary, although the contributing area was slightly larger and included an additional 414 acres of land. Land use composition was also very similar, with total land use for the drainage area of WBID 1633B being nearly 86 percent developed (72 percent high-intensity developed).

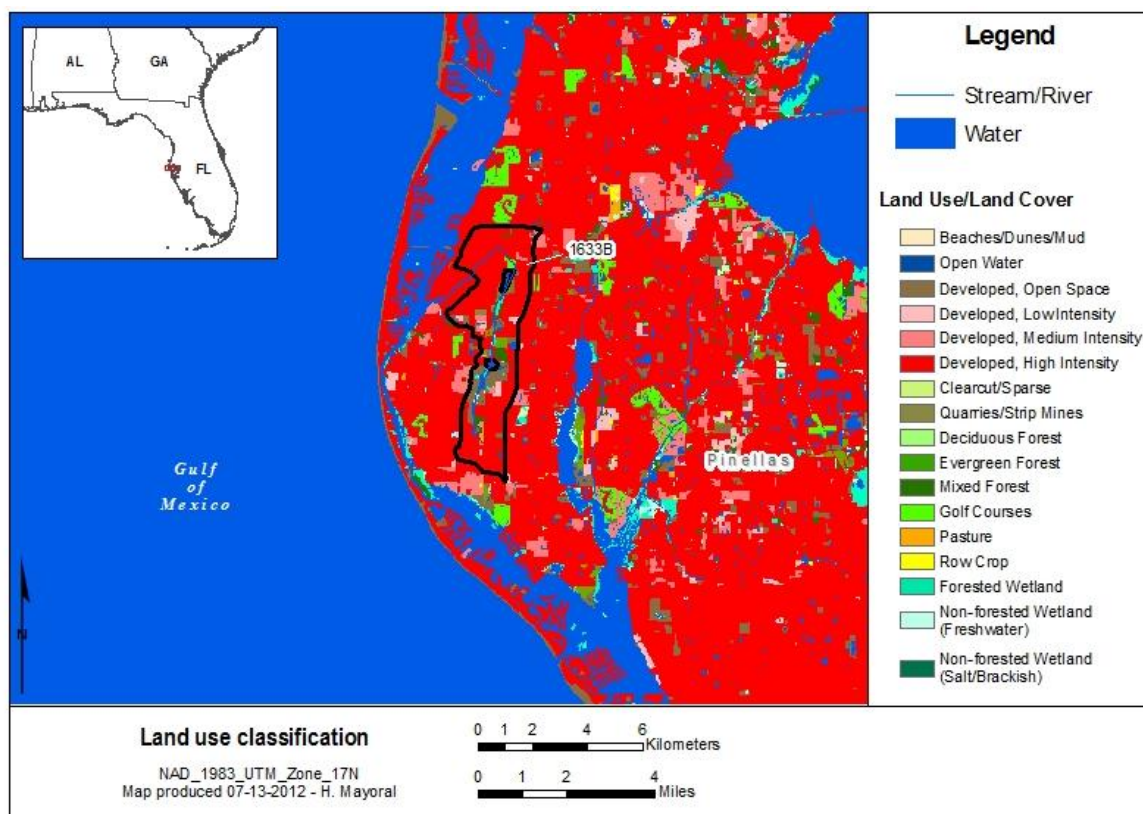


Figure 3.1 Land use for the impaired WBID in the McKay Creek watershed

Table 3.1 Land use distribution for WBID 1633B in the McKay Creek basin

Land Use Classification	WBID 1633B	
	Acres	%
Evergreen Forest	91	2%
Deciduous Forest	0	0%
Mixed Forest	105	3%
Forested Wetland	58	1%
Non-Forested Wetland (Freshwater)	19	0%
Open Water	155	4%
Pasture	25	1%
Row Crop	0	0%
Clear cut Sparse	0	0%
Quarries Strip mines	20	0%

Land Use Classification	WBID 1633B	
	Acres	%
Utility Swaths	0	0%
Developed, Open Space	461	11%
Developed, Low intensity	85	2%
Developed, Medium intensity	94	2%
Developed, High intensity	3,028	73%
Beaches/Dunes/Mud	0	0%
Golf Courses	31	1%
<b>Totals</b>	<b>4,172</b>	<b>100%</b>

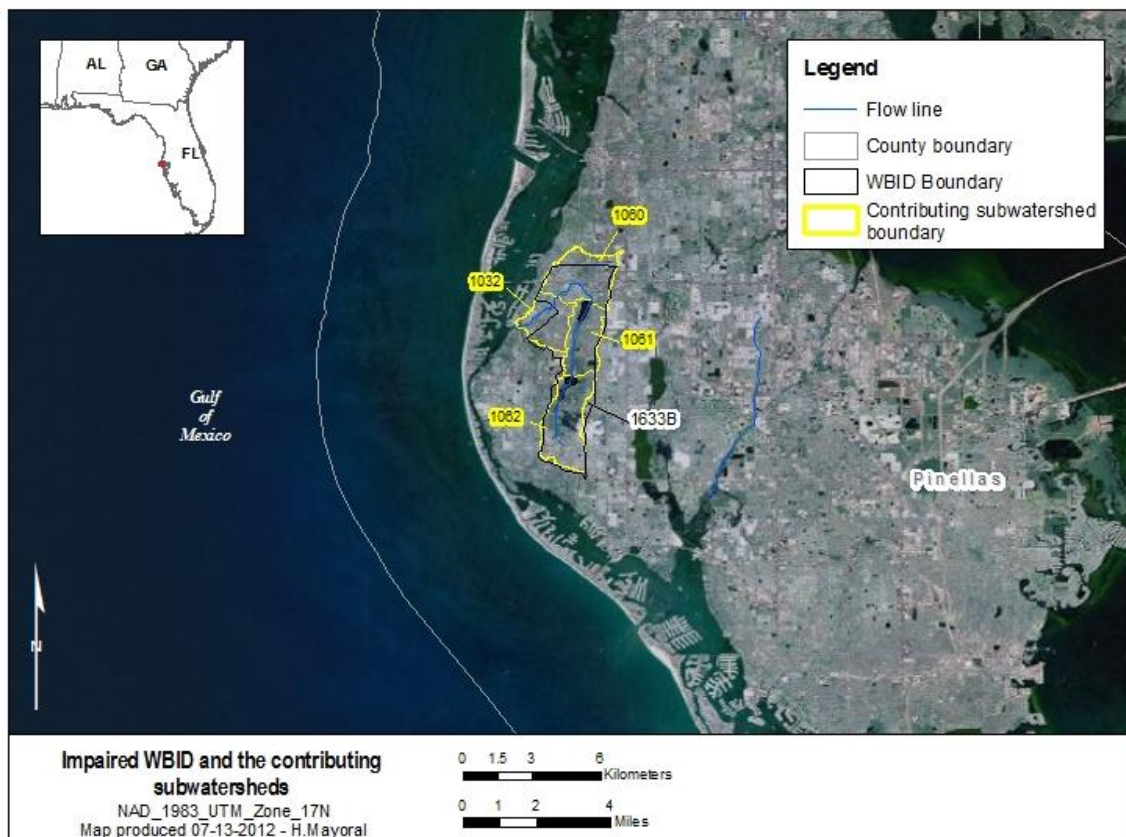


Figure 3.2 Aerial photograph illustrating contributing subwatersheds and impaired WBID boundaries

Table 3.2 Land use distribution for contributing subwatersheds in the McKay Creek watershed

Land Use Classification	Contributing subwatersheds	
	Acres	%
Evergreen Forest	91	2%
Deciduous Forest	0	0%
Mixed Forest	97	2%
Forested Wetland	58	1%
Non-Forested Wetland (Freshwater)	18	0%
Open Water	225	5%
Pasture	23	1%
Row Crop	0	0%
Clear cut Sparse	0	0%
Quarries Strip mines	20	0%
Utility Swaths	0	0%
Developed, Open Space	453	10%
Developed, Low intensity	96	2%
Developed, Medium intensity	93	2%
Developed, High intensity	3,312	72%
Beaches/Dunes/Mud	0	0%
Golf Courses	100	2%
Totals	4,586	100%

#### 4.0 WATER QUALITY STANDARDS/TMDL TARGETS

The TMDL reduction scenarios will be done to achieve Florida's dissolved oxygen concentration of 5 mg/L and insure balanced flora and fauna within McKay Creek or establish the TMDL to be consistent with a natural condition if the dissolved oxygen standard cannot be achieved.

The waterbody in the McKay Creek WBID is Class III Freshwater with a designated use of Recreation, Propagation and Maintenance of a Healthy, Well-Balanced Population of Fish and Wildlife. Designated use classifications are described in Florida's water quality standards. See Section 62-302.400, F.A.C. Water quality criteria for protection of all classes of waters are established in Section 62-302.530, F.A.C. Individual criteria should be considered in conjunction with other provisions in water quality standards, including Section 62-302.500 F.A.C., which established minimum criteria that apply to all waters unless alternative criteria are specified. Section 62-302.530, F.A.C. The WBID addressed in this report was listed due to both elevated concentrations of chlorophyll a, dissolved oxygen and/or elevated nitrogen and phosphorus concentrations. While FDEP does not have a streams water quality standard specifically for chlorophyll a, elevated levels of chlorophyll a are frequently associated with a violation of the narrative nutrient standard, which is described below.

#### **4.1 Nutrients Criteria**

The designated use of Class III waters is recreation, propagation and maintenance of a healthy, well-balanced population of fish and wildlife. In 1979, FDEP adopted a narrative criterion for nutrients. FDEP recently adopted numeric nutrient criteria for many Class III waters in the state, including streams, which numerically interprets part of the state narrative criterion for nutrients. While those criteria have been submitted to EPA for review pursuant to section 303(c) of the CWA, EPA has not completed that review. Therefore, for streams in Florida, the applicable nutrient water quality standard for CWA purposes remains the Class III narrative criterion.

As set out more fully below, should any new or revised state criteria for nutrients in streams in Florida become applicable for CWA purposes before this proposed TMDL is established, EPA will consider the impact of such criteria on the target selected for this TMDL.

Also, in November 2010, EPA promulgated numeric nutrient criteria for Class III inland waters in Florida, including streams. On February 18, 2012, the streams criteria were invalidated by the U.S. District Court for the Northern District of Florida and remanded back to EPA. Should a federally promulgated criterion become effective for CWA purposes before this proposed TMDL is established, EPA will consider the impact of such criteria on the target selected for this TMDL.

##### **4.1.1 Narrative Nutrient Criteria**

Florida's narrative nutrient criteria provide:

The discharge of nutrients shall continue to be limited as needed to prevent violations of other standards contained in this chapter. Man induced nutrient enrichment (total nitrogen and total phosphorus) shall be considered degradation in relation to the provisions of Sections 62-302.300, 62-302.700, and 62-4.242. Section 62-302.530(48)(a), F.A.C.

In no case shall nutrient concentrations of a body of water be altered so as to cause an imbalance in natural populations of aquatic flora or fauna. Section 62-302.530(48)(b), F.A.C.

Chlorophyll and DO levels are often used to indicate whether nutrients are present in excessive amounts. The target for this TMDL is based on levels of nutrients necessary to prevent violations of Florida's DO criterion, set out below.

#### 4.1.2 Inland Nutrients Criteria

Florida's recently adopted numeric nutrient criteria interprets the narrative water quality criterion for nutrients in paragraph 62-302.530(48)(b), F.A.C. See section 62-302.531(2). The Florida rule provides that the narrative water quality criteria for nutrients in paragraph 62-302.530(47)(a), F.A.C., continues to apply to all Class III waters. See section 62-302.531(1).

Florida's recently adopted rule applies to streams, including (WBID 1633B). For streams that do not have a site specific criteria, Florida's rule provides for biological information to be considered together with nutrient thresholds to determine whether a waterbody is attaining 62-302.531(2)(c), F.A.C. The rule provides that the nutrient criteria are attained in a stream segment where information on chlorophyll a levels, algal mats or blooms, nuisance macrophyte growth, and changes in algal species composition indicates there are no imbalances in flora and either the average score of at least two temporally independent SCIs performed at representative locations and times is 40 or higher, with neither of the two most recent SCI scores less than 35, or the nutrient thresholds set forth in Table 1 below are achieved. See section 62-302.531(2)(c).

Florida's rule provides that numeric nutrient criteria are expressed as a geometric mean, and concentrations are not to be exceeded more than once in any three calendar year period. Section 62-302.200 (25)(e), F.A.C.

Should FDEP's numeric nutrient criteria for streams become an applicable water quality standard for CWA purposes before this TMDL is established, EPA will consider the nutrient target necessary to attain section 62-302.531(2)(c), F.A.C. EPA will compare that target with the target necessary to attain paragraph 62-302.530(47)(a), F.A.C., to determine which target is more stringent.

Table 4.1 Inland numeric nutrient criteria

Nutrient Watershed Region	Total Phosphorus Nutrient Threshold	Total Nitrogen Nutrient Threshold
Panhandle West	0.06 mg/L	0.67 mg/L
Panhandle East	0.18 mg/L	1.03 mg/L
North Central	0.30 mg/L	1.87 mg/L
Peninsular	0.12 mg/L	1.54 mg/L
West Central	0.49 mg/L	1.65 mg/L
South Florida	No numeric nutrient threshold. The narrative criterion in paragraph 62-302.530(47)(b), F.A.C., applies.	No numeric nutrient threshold. The narrative criterion in paragraph 62-302.530(47)(b), F.A.C., applies.

## **4.2 Dissolved Oxygen Criteria**

Numeric criteria for DO are expressed in terms of minimum and daily average concentrations. Section 62-302(30), F.A.C., sets out the water quality criterion for the protection of Class III freshwater waters as:

Shall not be less than 5.0 mg/l. Normal daily and seasonal fluctuations above these levels shall be maintained.

## **4.3 Natural Conditions**

In addition to the standards for nutrients, DO and BOD described above, Florida's standards include provisions that address waterbodies which do not meet the standards due to natural background conditions.

Florida's water quality standards provide a definition of natural background:

"Natural Background" shall mean the condition of waters in the absence of man-induced alterations based on the best scientific information available to the Department. The establishment of natural background for an altered waterbody may be based upon a similar unaltered waterbody or on historical pre-alteration data. 62-302.200(15), FAC.

Florida's water quality standards also provide that:

Pollution which causes or contributes to new violations of water quality standards or to continuation of existing violations is harmful to the waters of this State and shall not be allowed. Waters having water quality below the criteria established for them shall be protected and enhanced. However, the Department shall not strive to abate natural conditions. 62-302.300(15) FAC

## **4.4 Biochemical Oxygen Demand Criteria**

Biochemical Oxygen Demand (BOD) shall not be increased to exceed values which would cause dissolved oxygen to be depressed below the limit established for each class and, in no case, shall it be great enough to produce nuisance conditions. [FAC 62-302.530 (11)]

The waterbody addressed in this report is a Class III water having a designated use of Recreation, Propagation and Maintenance of a Healthy, Well-Balanced Population of Fish and Wildlife. Designated use classifications are described in Florida's water quality standards in Section 62-302.400, FAC. Water quality criteria for protection of all classes of waters are established in Section 62-302.530, FAC. Individual criteria should be considered in conjunction with other provisions in water quality standards, including Section 62-302.500 FAC, which established minimum criteria that apply to all waters unless alternative criteria are specified Section 62-302.530, FAC. In addition, unless otherwise stated, all criteria express the maximum not to be exceeded at any time. The specific criteria addressed in this TMDL document are provided in the following sections.

## 5.0 WATER QUALITY ASSESSMENT

The WBID addressed in this report was listed as not attaining their designated use on Florida's 2009 303(d) list for dissolved oxygen and nutrients. To determine impairment, an assessment of available data was conducted. The source for current ambient monitoring data was the Impaired Waters Rule (IWR) data Run 44, using data ranging January 1, 2002 to December 31, 2010. The IWR database contains data from various sources within the state of Florida, including the WMDs and counties.

### 5.1 Water Quality Data

A complete list of water quality monitoring stations in WBID 1633B are located in Table 5.1, and an analysis of water quality data is documented in Table 5.2. Figure 5.1 shows the locations of the water quality monitoring stations within the WBID. Water quality data for the WBID can be found below in Figure 5.2 through Figure 5.6, with the data from all water quality stations compiled in each figure.

#### 5.1.1 Dissolved Oxygen

There are several factors that affect the concentration of dissolved oxygen (DO) in a waterbody, and natural DO levels are a function of water temperature, water depth and velocity, salinity and relative contributions from groundwater. Oxygen can be introduced by wind, diffusion, photosynthesis, and additions of higher DO water (e.g. from tributaries). DO concentrations can be lowered by processes that use up oxygen from the water, such as respiration and decomposition, and can be lowered through additions of water with lower DO (e.g. swamp or groundwater). Decomposition of organic matter, such as dead plants and animals, also consume DO. The dissolved oxygen minimum concentration was 0.59 mg/L, and the maximum concentration was 11.85 mg/L. The mean concentration was 5.73 mg/L.

#### 5.1.2 Biochemical Oxygen Demand

BOD is a measure of the amount of oxygen used by bacteria as they stabilize organic matter. The process can be accelerated when there is an overabundance of nutrients, increasing the aerobic bacterial activity in a waterbody. In turn, the levels of DO can become depleted, eliminating oxygen essential for biotic survival, and potentially causing extensive fish kills. Additionally, BOD is used as an indicator to determine the presence and magnitude of organic pollution from sources such as septic tank leakage, fertilizer runoff, and wastewater effluent. The mean BOD concentration for WBID 1633B was 4.62 mg/L. The maximum BOD concentration was 8.40 mg/L and the minimum concentration was 0.84 mg/L.

#### 5.1.3 Nutrients

Excessive nutrients in a waterbody can lead to overgrowth of algae and other aquatic plants such as phytoplankton, periphyton and macrophytes. This process can deplete oxygen in the water, adversely affecting aquatic life and potentially restricting recreational uses such as fishing and boating. For the nutrient assessment the monitoring data for total nitrogen, total phosphorus and chlorophyll a are presented. Narrative nutrient criteria are used as the standards for estuarine

water bodies, while numeric standards have been developed for freshwater water bodies. The purpose of the nutrient assessment is to present the range, variability and average conditions for the WBID.

#### **5.1.3.1 Total Nitrogen**

Total Nitrogen (TN) is comprised of nitrate (NO<sub>3</sub>), nitrite (NO<sub>2</sub>), organic nitrogen and ammonia nitrogen (NH<sub>4</sub>). Though nitrogen is a necessary nutrient required for the growth of most plants and animals, not all forms are readily used or metabolized. Increased levels of organic nitrogen can occur from the decomposition of aquatic life or from sewage, while inorganic forms are generally from erosion and fertilizers. Nitrates are components of industrial fertilizers, yet can also be naturally present in soil, and are converted to nitrite by microorganisms in the environment. Surface runoff from agricultural lands can increase the natural presence of nitrates in the environment and can lead to eutrophication. Usually, the eutrophication process is observed as a change in the structure of the algal community and includes severe algal blooms that may cover large areas for extended periods. Large algal blooms are generally followed by depletion in DO concentrations as a result of algal decomposition. The total nitrogen minimum concentration was 0.42 mg/L, and the maximum concentration was 4.06 mg/L. The mean total nitrogen concentration in WBID 1633B was 0.98 mg/L.

#### **5.1.3.2 Total Phosphorus**

In natural waters, total phosphorus exists in either soluble or particulate forms. Dissolved phosphorus includes inorganic and organic forms, while particulate phosphorus is made up of living and dead plankton, and adsorbed, amorphous, and precipitated forms. Inorganic forms of phosphorus include orthophosphate and polyphosphates, though polyphosphates are unstable and convert to orthophosphate over time. Orthophosphate is both stable and reactive, making it the form most used by plants. Excessive phosphorus can lead to overgrowth of algae and aquatic plants, the decomposition of which depletes oxygen in the water. The total phosphorus minimum concentration was 0.01 mg/L, and the maximum concentration was 0.54 mg/L. The mean total phosphorus concentration in WBID 1633B was 0.11 mg/L.

#### **5.1.3.3 Chlorophyll-a**

Chlorophyll is the green pigment in plants that allows them to create energy from light. In a water sample, chlorophyll is indicative of the presence of algae, and chlorophyll-*a* is a measure of the active portion of total chlorophyll. Corrected chlorophyll refers to chlorophyll-*a* measurements that are corrected for the presence of pheophytin, a natural degradation product of chlorophyll that can interfere with analysis because it has an absorption peak in the same spectral region. It is used as a proxy indicator of water quality because of its predictable response to nutrient availability. Increases in nutrients can potentially lead to blooms in phytoplankton biomass, affecting water quality and ecosystem health. The corrected chlorophyll a maximum concentration was 47.0 µg/L, and the mean concentration was 7.61 µg/L.

Table 5.1 Water quality stations located in WBID 1633B

WBID	Station Number
1633B	21FLGW 35437
	21FLPDEM27-03
	21FLPDEM27-09
	21FLPDEM27-10
	21FLPDEMAMB 27-3
	21FLTPA 27525378248329
	21FLTPA 27543408248589
	21FLTPA 27544608248480
	21FLTPA 27545608248150
	21FLTPA 27550008248318

Table 5.2 Water quality data for WBID 1633B

Parameter	Stats	WBID 1633B
BOD, 5 Day, 20°C (mg/L)	# of obs	56
	min	0.84
	max	8.40
	mean	4.62
	Geomean	3.66
DO, Analysis by Probe (mg/L)	# of obs	277
	min	0.59
	max	11.85
	mean	5.73
	Geomean	5.07
Nitrogen, Total (mg/L as N)	# of obs	178
	min	0.42
	max	4.06

Parameter	Stats	WBID 1633B
	mean	0.98
	Geomean	0.93
Phosphorus, Total (mg/L as P)	# of obs	173
	min	0.01
	max	0.54
	mean	0.11
	Geomean	0.07
Chlorophyll-A-corrected (µg/L)	# of obs	181
	min	1.00
	max	47.00
	mean	7.61
	Geomean	4.40

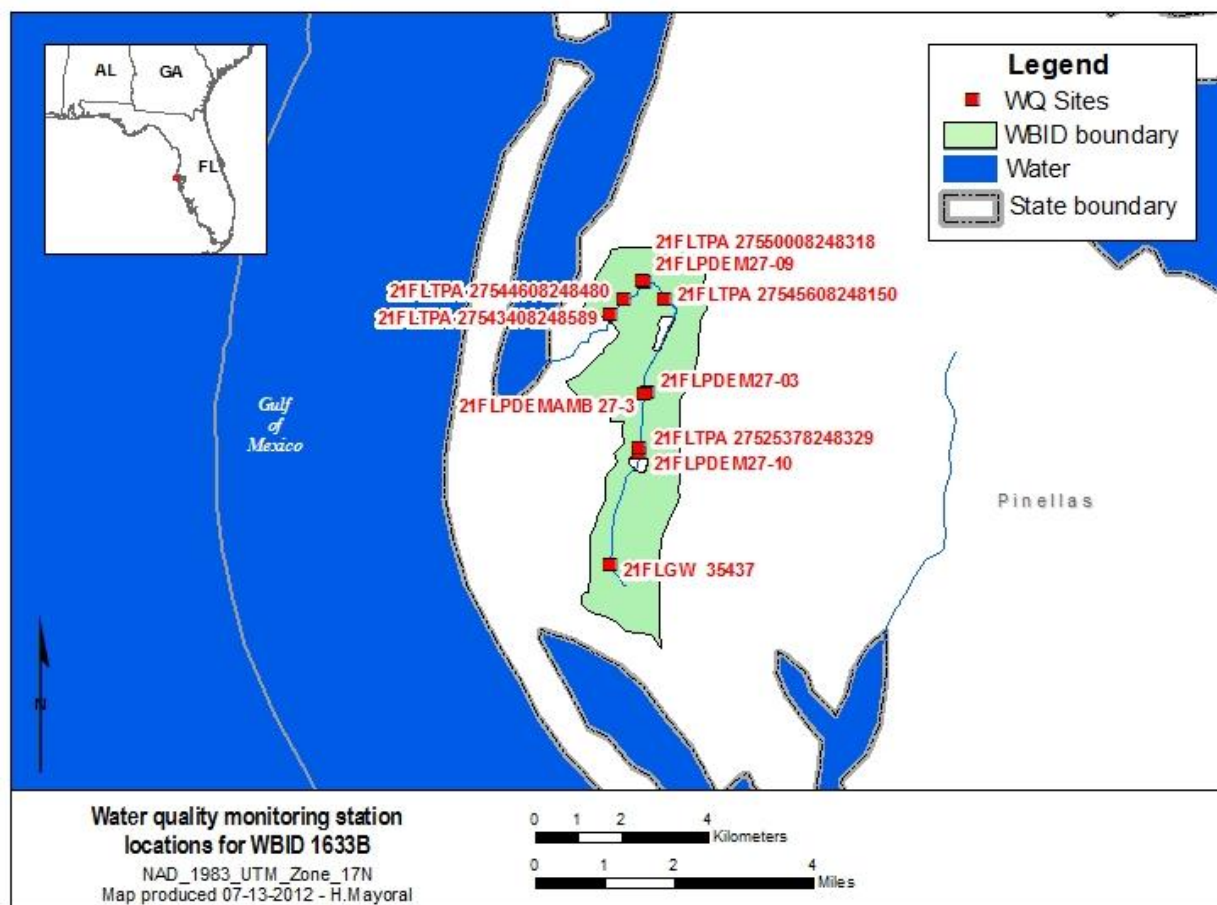


Figure 5.1 Water quality monitoring station locations for impaired WBID 1633B in the McKay Creek watershed

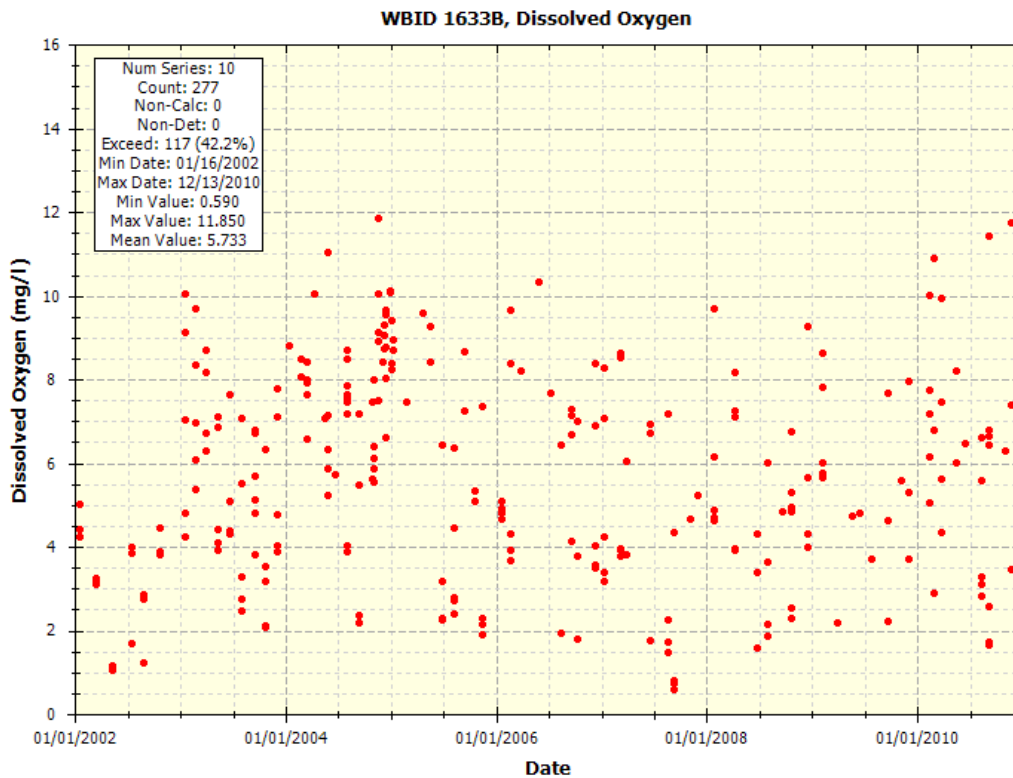


Figure 5.2 Dissolved oxygen concentrations for WBID 1633B

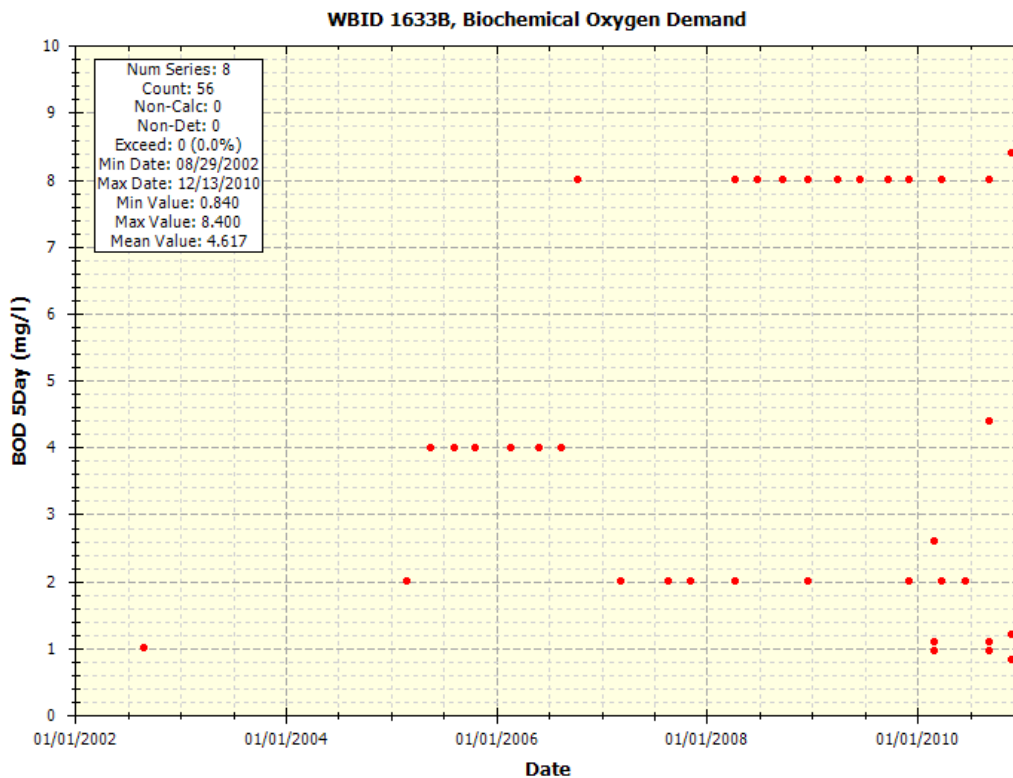


Figure 5.3 Biochemical oxygen demand concentrations for WBID 1633B

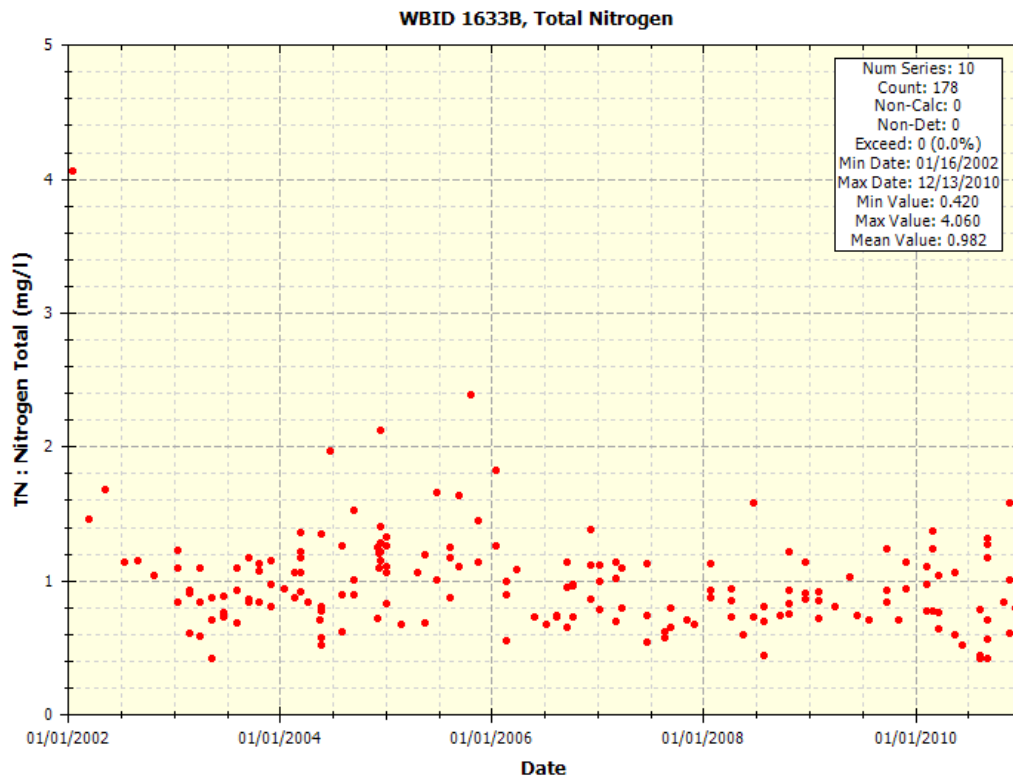


Figure 5.4 Total nitrogen concentrations for WBID 1633B

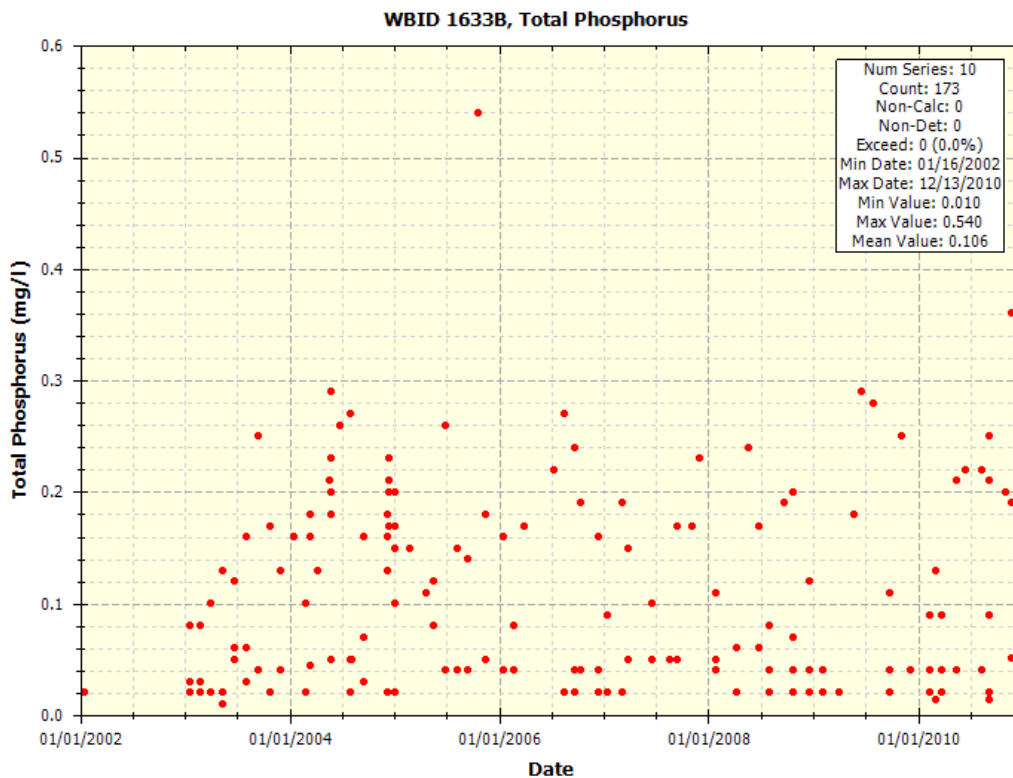


Figure 5.5 Total phosphorus concentrations for WBID 1633B

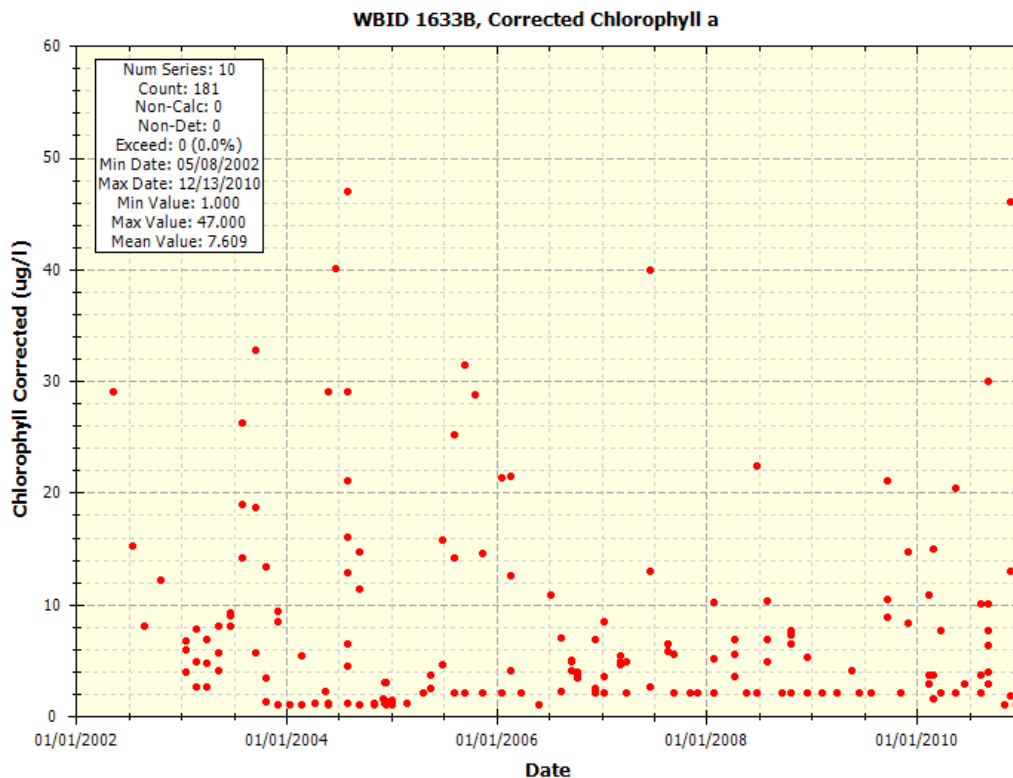


Figure 5.6 Corrected chlorophyll a concentrations for WBID 1633B

## 6.0 SOURCE AND LOAD ASSESSMENT

An important part of the TMDL analysis is the identification of source categories, source subcategories, or individual sources of pollutants in the watershed and the amount of loading contributed by each of these sources. Sources are broadly classified as either point or nonpoint sources. Nutrients can enter surface waters from both point and nonpoint sources.

### 6.1 Point Sources

A point source is defined as a discernable, confined, and discrete conveyance from which pollutants are or may be discharged to surface waters. Point source discharges of industrial wastewater and treated sanitary wastewater must be authorized by National Pollutant Discharge Elimination System (NPDES) permits. NPDES permitted discharges include continuous discharges such as wastewater treatment facilities as well as some stormwater driven sources such as municipal separate stormwater sewer systems (MS4s), certain industrial facilities, and construction sites over one acre.

### 6.1.1 Wastewater/Industrial Permitted Facilities

A TMDL wasteload allocation (WLA) is given to wastewater and industrial NPDES permitted facilities discharging to surface waters within an impaired watershed. There are no NPDES-permitted facilities in WBID 1633B.

### 6.1.2 Stormwater Permitted Facilities/MS4s

MS4s are point sources also regulated by the NPDES program. According to 40 CFR 122.26(b)(8), an MS4 is “a conveyance or system of conveyances (including roads with drainage systems, municipal streets, catch basins, curbs, gutters, ditches, man-made channels, or storm drains):

- (i) Owned or operated by a State, city, town, borough, county, parish, district, association, or other public body (created by or pursuant to State law)...including special districts under State law such as a sewer district, flood control district or drainage district, or similar entity, or an Indian tribe or an authorized Indian tribal organization, or a designated and approved management agency under section 208 of the Clean Water Act that discharges into waters of the United States.
- (ii) Designed or used for collecting or conveying storm water;
- (iii) Which is not a combined sewer; and
- (iv) Which is not part of a Publicly Owned Treatment Works

MS4s may discharge nutrients and other pollutants to waterbodies in response to storm events. In 1990, USEPA developed rules establishing Phase I of the NPDES stormwater program, designed to prevent harmful pollutants from being washed by stormwater runoff into MS4s (or from being dumped directly into the MS4) and then discharged from the MS4 into local waterbodies. Phase I of the program required operators of “medium” and “large” MS4s (those generally serving populations of 100,000 or greater) to implement a stormwater management program as a means to control polluted discharges from MS4s. Approved stormwater management programs for medium and large MS4s are required to address a variety of water quality related issues including roadway runoff management, municipal owned operations, hazardous waste treatment, etc.

Phase II of the rule extends coverage of the NPDES stormwater program to certain “small” MS4s. Small MS4s are defined as any MS4 that is not a medium or large MS4 covered by Phase I of the NPDES stormwater program. Only a select subset of small MS4s, referred to as “regulated small MS4s”, requires an NPDES stormwater permit. Regulated small MS4s are defined as all small MS4s located in “urbanized areas” as defined by the Bureau of the Census, and those small MS4s located outside of “urbanized areas” that are designated by NPDES permitting authorities.

In October 2000, USEPA authorized FDEP to implement the NPDES stormwater program in all areas of Florida except Indian tribal lands. FDEP’s authority to administer the NPDES program is set forth in Section 403.0885, Florida Statutes (FS). The three major components of NPDES stormwater regulations are:

- MS4 permits that are issued to entities that own and operate master stormwater systems, primarily local governments. Permittees are required to implement comprehensive stormwater management programs designed to reduce the discharge of pollutants from the MS4 to the maximum extent practicable.
- Stormwater associated with industrial activities, which is regulated primarily by a multisector general permit that covers various types of industrial facilities. Regulated industrial facilities must obtain NPDES stormwater permit coverage and implement appropriate pollution prevention techniques to reduce contamination of stormwater.
- Construction activity general permits for projects that ultimately disturb one or more acres of land and which require the implementation of stormwater pollution prevention plans to provide for erosion and sediment control during construction.

Stormwater discharges conveyed through the storm sewer system covered by the permit are subject to the WLA of the TMDL. Any newly designated MS4s will also be required to achieve the percent reduction allocation presented in this TMDL. There are two MS4 permits. One Phase I C MS4 associated with the impaired WBID, for Pinellas County (FLS000005), which also falls under the District VII Florida Department of Transportation permit. There are three Phase I MS4s which fall under the Pinellas County permit FLS000005 as co-permittees for the Cities of Belleair Bluffs, Clearwater, and Largo. The City of St. Petersburg holds its own separate Phase I MS4 permit (FLS000007).

## **6.2 Nonpoint Sources**

Nonpoint sources of pollution are diffuse sources that cannot be identified as entering a waterbody through a discrete conveyance at a single location. For nutrients, these sources include runoff of agricultural fields, golf courses, and lawns, septic tanks, and residential developments outside of MS4 areas. Nonpoint source pollution generally involves a buildup of pollutants on the land surface that wash off during rain events and as such, represent contributions from diffuse sources, rather than from a defined outlet. Potential nonpoint sources are commonly identified, and their loads estimated, based on land cover data. Most methods calculate nonpoint source loadings as the product of the water quality concentration and runoff water volume associated with certain land use practices. The mean concentration of pollutants in the runoff from a storm event is known as the event mean concentration. Figure 3.1 provides a map of the land use, while Table 3.1 lists the land use distribution in the WBID.

The following sections are organized by land use. Each section provides a description of the land use, the typical sources of nutrient loading (if applicable), and typical total nitrogen and total phosphorus event mean concentrations.

### **6.2.1 Urban Areas**

Urban areas include land uses such as residential, industrial, extractive and commercial. Land uses in this category typically have somewhat high total nitrogen event mean concentrations and average total phosphorus event mean concentrations. Nutrient loading from MS4 and non-MS4 urban areas is attributable to multiple sources including stormwater runoff, leaks and overflows

from sanitary sewer systems, illicit discharges of sanitary waste, runoff from improper disposal of waste materials, leaking septic systems, and domestic animals.

In 1982, Florida became the first state in the country to implement statewide regulations to address the issue of nonpoint source pollution by requiring new development and redevelopment to treat stormwater before it is discharged. The Stormwater Rule, as outlined in Chapter 403 FS, was established as a technology-based program that relies upon the implementation of Best Management Practices (BMPs) that are designed to achieve a specific level of treatment (i.e., performance standards) as set forth in Chapter 62-40, FAC.

Florida's stormwater program is unique in having a performance standard for older stormwater systems that were built before the implementation of the Stormwater Rule in 1982. This rule states: "the pollutant loading from older stormwater management systems shall be reduced as needed to restore or maintain the beneficial uses of water." [FAC 62-40-.432(2)(c)]

Nonstructural and structural BMPs are an integral part of the State's stormwater programs. Nonstructural BMPs, often referred to as "source controls", are those that can be used to prevent the generation of nonpoint source pollutants or to limit their transport off-site. Typical nonstructural BMPs include public education, land use management, preservation of wetlands and floodplains, and minimization of impervious surfaces. Technology-based structural BMPs are used to mitigate the increased stormwater peak discharge rate, volume, and pollutant loadings that accompany urbanization.

Urban, residential, and commercial developments are often a significant nonpoint source of nutrients and oxygen-demanding substances. In WBID 1633B, 86 percent of the contributing land use is developed, a majority consisting of high intensity developments; indicating that urban land uses are likely a significant cause of the impairment.

### ***Onsite Sewage Treatment and Disposal Systems (Septic Tanks)***

As stated above leaking septic tanks or onsite sewage treatment and disposal systems (OSTDs) can contribute to nutrient loading in urban areas. Water from OSTDs is typically released to the ground through on-site, subsurface drain fields or boreholes that allow the water from the tank to percolate (usually into the surficial aquifers) and either transpire to the atmosphere through surface vegetation or add to the flow of shallow ground water. When properly sited, designed, constructed, maintained, and operated, OSTDs are a safe means of disposing of domestic waste. The effluent from a well-functioning OSTD receives natural biological treatment in the soil and is comparable to secondarily treated wastewater from a sewage treatment plant. When not functioning properly, OSTDs can be a source of nutrients, pathogens, and other pollutants to both ground water and surface water.

The State of Florida Department of Health publishes data on new septic tank installations and the number of septic tank repair permits issued for each county in Florida. Table 6.1 summarizes the cumulative number of septic systems installed in Pinellas County since the 1970 census and the total number of repair permits issued for the ten years between 1999-2000 and 2009-2010 (FDOH 2009). The data do not reflect septic tanks removed from service. Leaking septic systems could be a relevant source of organic and nutrient loading in the watershed.

Table 6.1 County estimates of Septic Tanks and Repair Permits

County	Number of Septic Tanks (1970-2008)	Number of Repair Permits Issued (2000-2010)
Pinellas	23,869	3,015

Note: Source: <http://www.doh.state.fl.us/environment/ostds/statistics/ostdsstatistics.htm>

### 6.2.2 Pastures

Pastures include cropland and improved and unimproved pasturelands, such as non-tilled grasses woodland pastures, feeding operations, nurseries and vineyards; as well as specialty farms. Agricultural activities, including runoff of fertilizers or animal wastes from pasture and cropland and direct animal access to streams, can generate nutrient loading to streams. The highest total nitrogen and total phosphorus event mean concentrations are associated with agricultural land uses. Pastures account for less than one percent of the total land use within WBID 1633B and therefore are not likely a source of excessive nutrients.

### 6.2.3 Clear cut/Sparse

The clear cut/sparse land use classification includes recent clear cuts, areas of sparse vegetation or herbaceous dry prairie, shrub and brushland, other early successional areas, and mixed rangeland. Event mean concentrations for clear cut/sparse can be relatively low for total nitrogen and total phosphorus. There are no areas of clear cut/sparse land use within WBID 1633B.

### 6.2.4 Forests

Upland forests include flatwoods, oak, various types of hardwoods, conifers and tree plantations. Wildlife, located within forested areas, deposit their feces onto land surfaces where it can be transported to nearby streams during storm events. Generally, the pollutant load from wildlife is assumed to represent background concentrations. Event mean concentrations for upland forests are low for both total nitrogen and total phosphorus. Combined forested land use accounts for the 4 percent of the contributing land use to WBID 1633B.

### 6.2.5 Water and Wetlands

Water and Wetlands often have very low nutrient loadings, although decaying organic matter in wetlands can contribute to high organic nutrient concentrations. Open water accounts for 5 percent of total land use in WBID 1633B, while both forested and non-forested wetlands combined account for 2 percent of total land use.

### 6.2.6 Quarries/Strip mines

Land use classification includes quarries, strip mines, exposed rock and soil, fill areas, reclaimed lands, and holding ponds. Event mean concentrations for some barren lands tend to be higher in total nitrogen. Less than one percent of total land use is made up of quarries/strip mines in WBID 1633B.

## 7.0 ANALYTICAL APPROACH

In the development of a TMDL there needs to be a method for relating current loadings to the observed water quality problem. This relationship could be: statistical (regression for a cause and effect relationship), empirical (based on observations not necessarily from the waterbody in question) or mechanistic (physically and/or stochastically based) that inherently relate cause and effect using physical and biological relationships.

Mechanistic models were used in the development of the McKay Creek TMDL to relate the physical and biological relationships. A dynamic watershed model was used to predict the quantity of water and pollutants associated with runoff from rain events. The watershed model was linked to a hydrodynamic model that simulated tidal influences in the river. Both models were linked to a water quality simulation model that integrated the loadings and flow from the watershed model with flow from the hydrodynamic model to predict the water quality in the receiving waterbodies.

The period of simulation that was considered in the development of this TMDL is January 1, 2002 to December 31, 2011. The models were used to predict time series for BOD, TN, TP, and DO. The models were calibrated to current conditions and were then used to predict improvements in water quality as function of reductions in loadings.

### 7.1 *Mechanistic Models*

#### 7.1.1 Loading Simulation Program C++ (LSPC)

LSPC is the Loading Simulation Program in C++, a watershed modeling system that includes streamlined Hydrologic Simulation Program Fortran (HSPF) algorithms for simulating hydrology, sediment, and general water quality overland as well as a simplified stream fate and transport model. LSPC is derived from the Mining Data Analysis System (MDAS), which was originally developed by USEPA Region 3 (under contract with Tetra Tech) and has been widely used for TMDLs. In 2003, the USEPA Region 4 contracted with Tetra Tech to refine, streamline, and produce user documentation for the model for public distribution. LSPC was developed to serve as the primary watershed model for the USEPA TMDL Modeling Toolbox. LSPC was used to simulate runoff (flow, biochemical oxygen demand, total nitrogen, total phosphorus and dissolved oxygen) from the land surface using a daily timestep for current and natural conditions. LSPC provided tributary flows and temperature to the EFDC estuary models and tributary water quality concentrations to WASP7 estuary models.

An LSPC model was utilized to estimate the nutrient loads within and discharged from the McKay Creek watershed. The LSPC model utilized the data inputs, including land use and weather data, from the larger Crystal Watershed model (USEPA 2012a and USEPA 2012b).

In order to evaluate the contributing sources to a waterbody and to represent the spatial variability of these sources within the watershed model, the contributing drainage area was represented by a series of sub-watersheds for each of the models. The sub-watersheds for the Crystal Watershed model were developed using the 12-digit hydrologic unit code (HUC12) watershed data layer and the Geological Survey (USGS) National Hydrograph Dataset (NHD).

The sub-watersheds were re-delineated at a smaller scale for the McKay Creek watershed model, which used the Pinellas County subwatershed delineations (Figure 7.1). Church Creek was included in the delineation because it drains into the estuary.

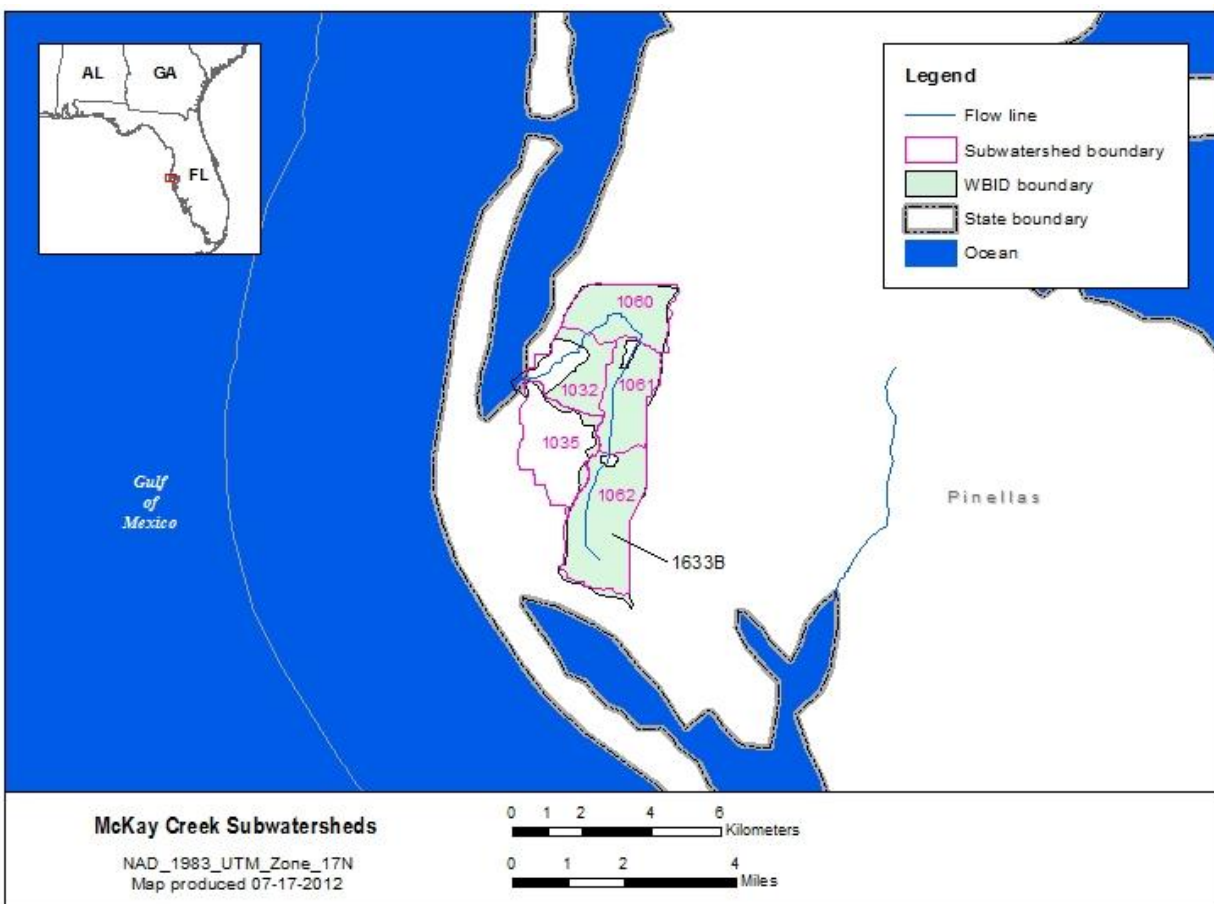


Figure 7.1 Location of McKay Creek LSPC subwatersheds

The LSPC model has a representative reach defined for each sub-watershed, and the main channel stem within each sub-watershed was used as the representative reach. The characteristics for each reach include the length and slope of the reach, the channel geometry and the connectivity between the sub-watersheds. Length and slope data for each reach was obtained using the USGS DEM and NHD data.

The attributes supplied for each reach were used to develop a function table (FTABLE) that describes the hydrology of the stream reach by defining the functional relationship between water depth, surface area, water volume, and outflow in the segment. The assumption of a fixed depth, area, volume, outflow relationship rules out cases where the flow reverses direction or where one reach influences another upstream of it in a time-dependent way. LSPC does not model the tidal flow in the low-lying estuaries, and therefore the main Crystal Watershed model was calibrated to non-tidally influenced USGS gages. The McKay Creek Watershed model was linked to the EFDC and WASP models to simulate the areas of the estuary that were tidally influenced.

The watershed model uses land use data as the basis for representing hydrology and nonpoint source loadings. The FDEP Level III Florida Land Use, specifically the Southwest Florida Water Management District (SWFWMD) 2004 dataset, was used to determine the land use representation. The National Landuse Coverage Dataset (NLCD) was used to develop the impervious land use representations.

The SWFWMD coverage utilized a variety of land use classes which were grouped and reclassified into 18 land use categories: beaches/dune/mud, open water, utility swaths, developed open space, developed low intensity, developed medium intensity, developed high intensity, clear-cut/sparse, quarries/strip mines, deciduous forest, evergreen forest, mixed forest, golf courses, pasture, row crop, forested wetland, non-forested wetland (salt/brackish), and non-forested wetland (freshwater). The LSPC model requires division of land uses in each sub-watershed into separate pervious and impervious land units. The NLCD 2006 percent impervious coverage was used to determine the percent impervious area associated with each land use category. Any impervious areas associated with utility swaths, developed open space, and developed low intensity, were grouped together and placed into a new land use category named *low intensity development impervious*. Impervious areas associated with medium intensity development and high intensity development were kept separate and placed into two new categories for *medium intensity development impervious* and *high intensity development impervious*, respectively. Finally, any impervious area not already accounted for in the three developed impervious categories, were grouped together into a fourth new category for all remaining impervious land use.

Soil data for the Florida watersheds was obtained from the Soil Survey Geographic Database (SSURGO). The database was produced and distributed by the Natural Resources Conservation Service (NRCS) - National Cartography and Geospatial Center (NCGC). The SSURGO data was used to determine the total area that each hydrologic soil group covered within each sub-watershed. The sub-watersheds were represented by the hydrologic soil group that had the highest percentage of coverage within the boundaries of the sub-watershed. There were four hydrologic soil groups which varied in their infiltrations rates and water storage capacity.

In the watershed models, nonpoint source loadings and hydrological conditions are dependent on weather conditions. Hourly data from weather stations within the boundaries of, or in close proximity to, the sub-watersheds were applied to the watershed model. A weather data forcing file was generated in ASCII format (\*.air) for each meteorological station used in the hydrological evaluations in LSPC. Each meteorological station file contained atmospheric data used in modeling the hydrological processes. These data included precipitation, air temperature, dew point temperature, wind speed, cloud cover, evaporation, and solar radiation. These data are used directly, or calculated from the observed data. The Crystal Watershed model weather stations only contained data through 2009. These stations were appended with data through 2011, and the LSPC model was evaluated through 2011.

The hydrodynamic calibration parameters from the larger Crystal Watershed model were used to populate the McKay Creek Watershed model. The Crystal Watershed model was calibrated to continuous flow USGS gages. No continuous measured flow data was located in the McKay Creek watershed, but several instantaneous measurements were taken in the freshwater portion of McKay Creek (Figure 7.2 and Figure 5.3). Additionally, the water quality parameters from

the larger Crystal Watershed model were used to populate the McKay Creek Watershed model. The Crystal Watershed model was calibrated to several water quality stations whose data was taken from IWR38. The McKay Creek watershed was calibrated to water quality data from IWR44, specifically to stations 21FLPDEM27-10 and 21FLDEPM27-09, which both contained large data records for all parameters of interest. LSPC water quality calibration results are presented in Figure 7.2 through Figure 7.15.

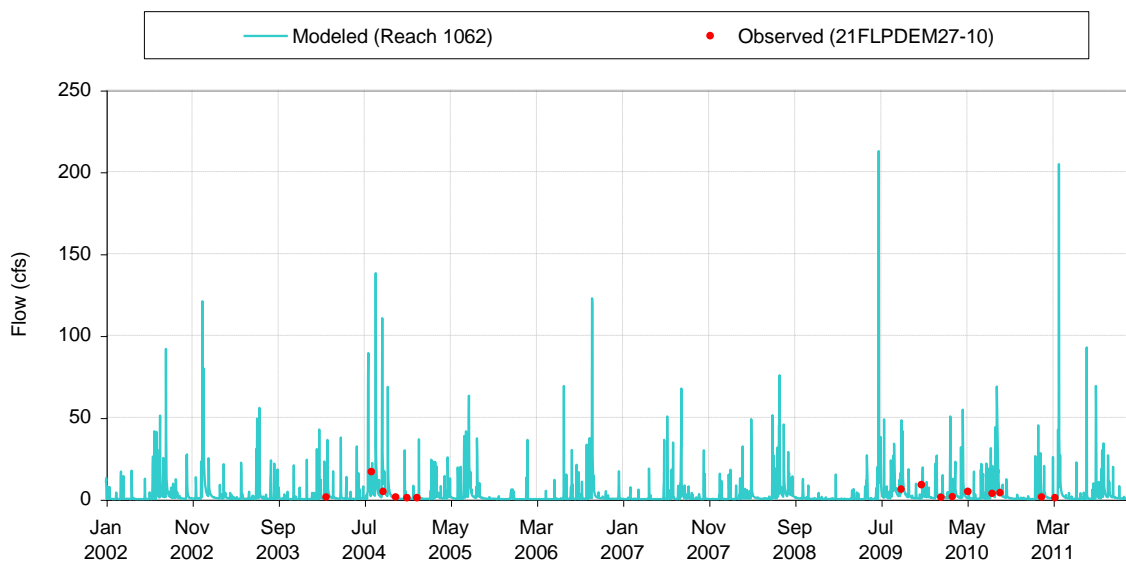


Figure 7.2 Modeled vs. Observed Flow (cfs) at 21FLPDEM27-10

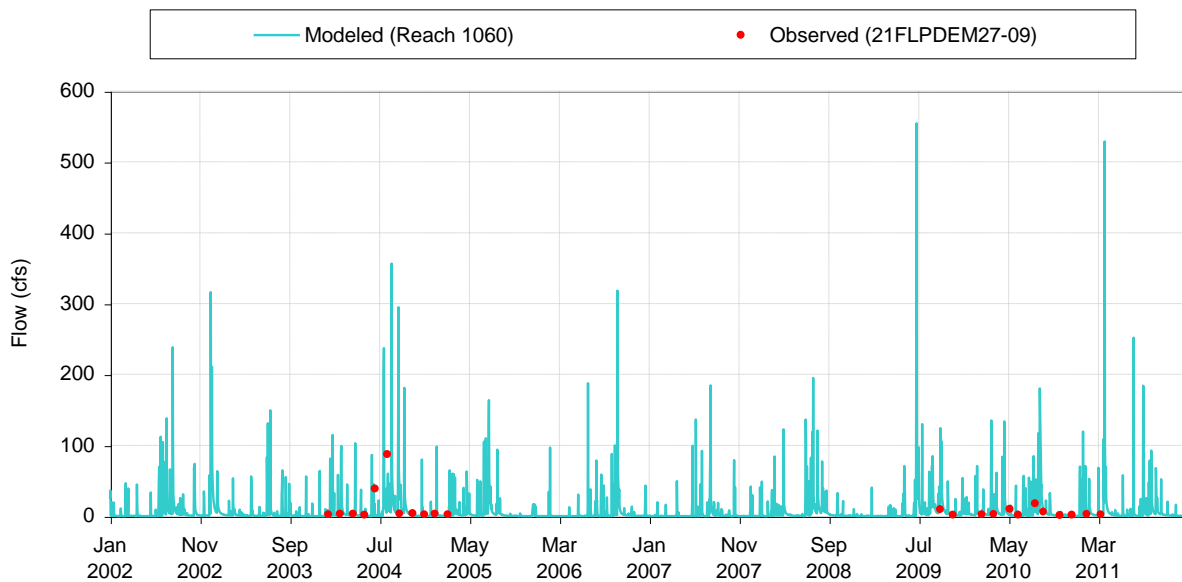


Figure 7.3 Modeled vs. Observed Flow (cfs) at 21FLPDEM27-09

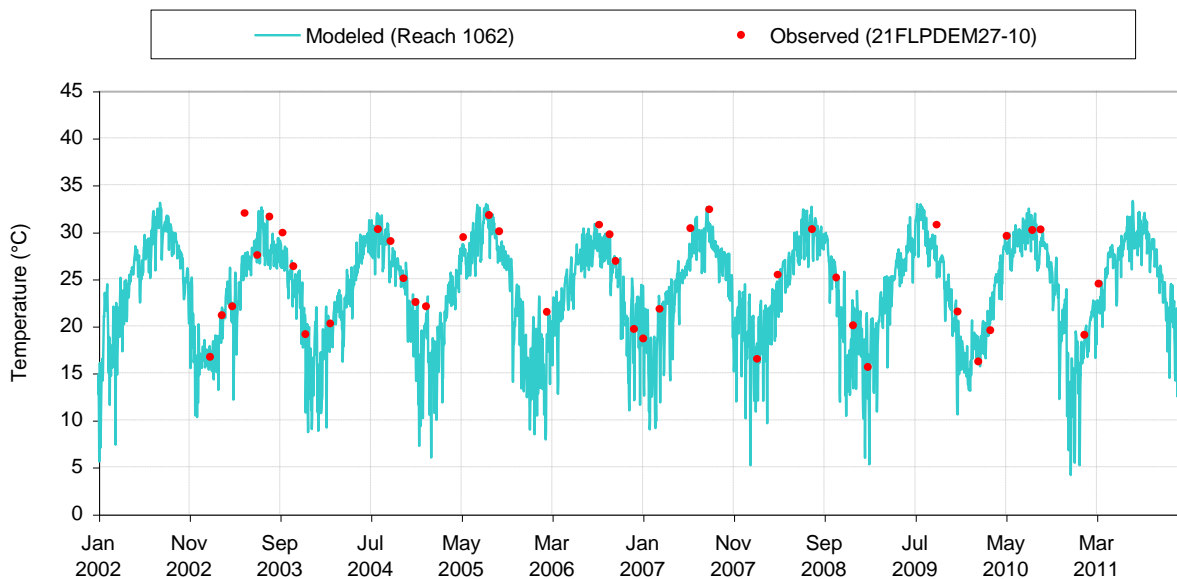


Figure 7.4 Modeled vs. Observed Temperature (°C) at 21FLPDEM27-10

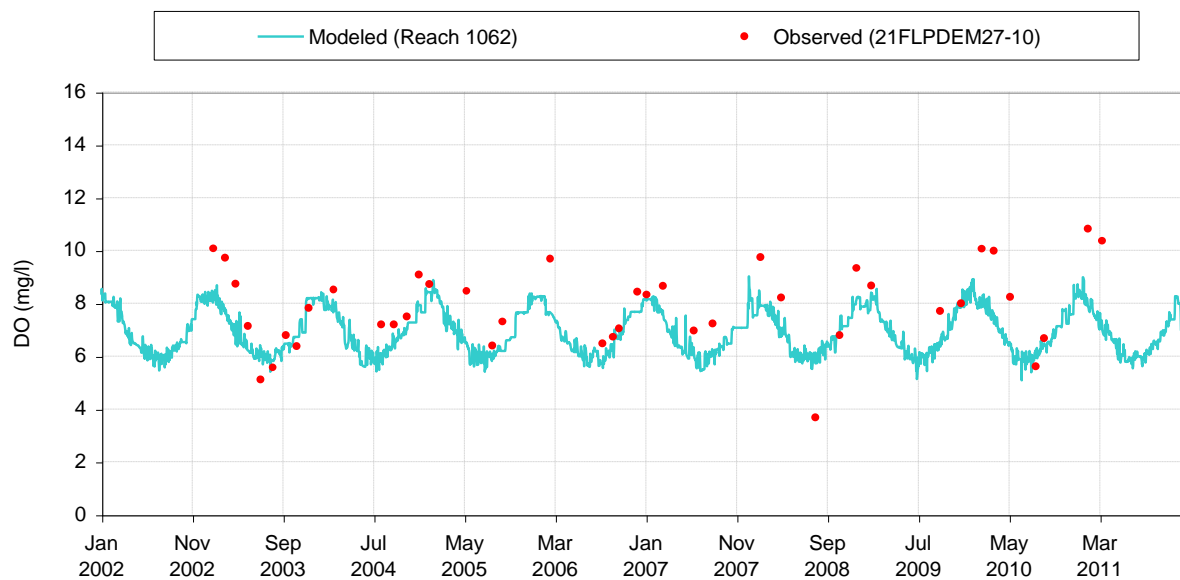


Figure 7.5 Modeled vs. Observed DO (mg/l) at 21FLPDEM27-10

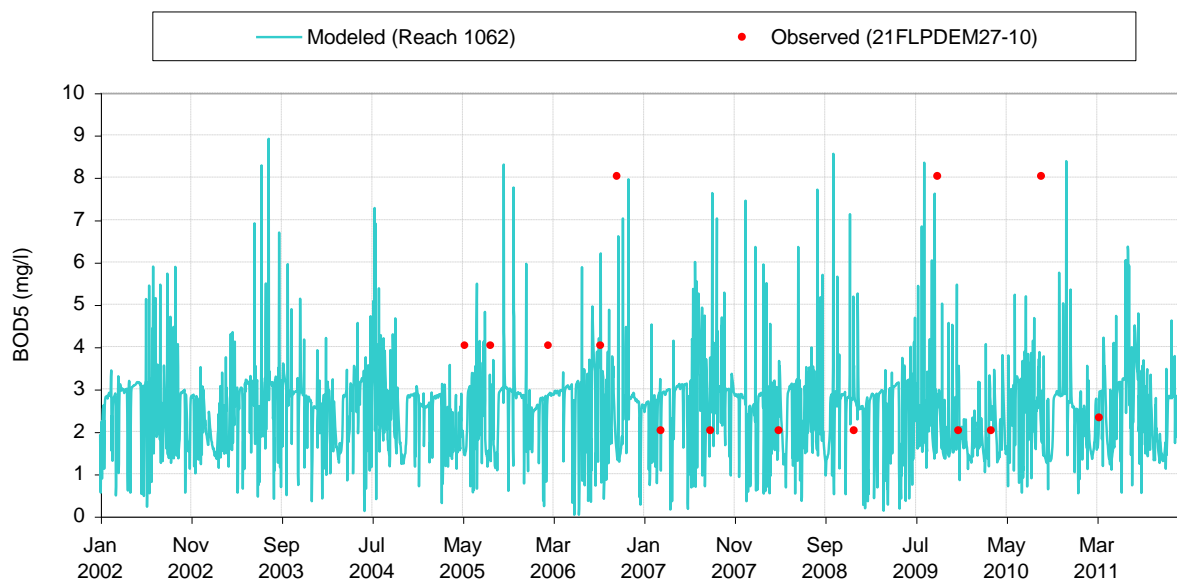


Figure 7.6 Modeled vs. Observed BOD<sub>5</sub> (mg/l) at 21FLPDEM27-10

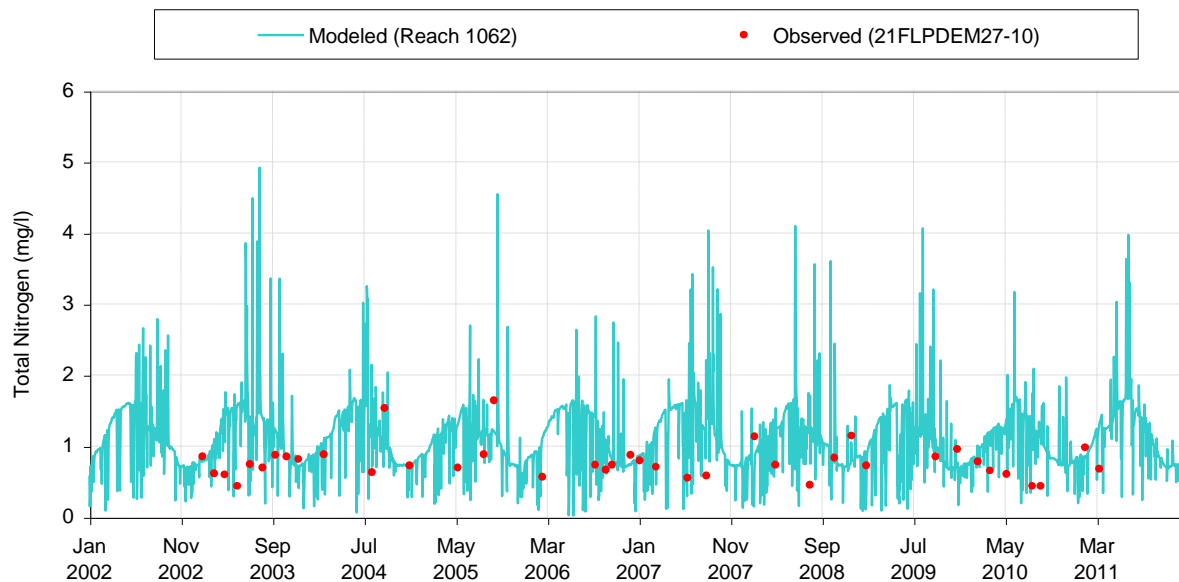


Figure 7.7 Modeled vs. Observed Total Nitrogen (mg/l) at 21FLPDEM27-10

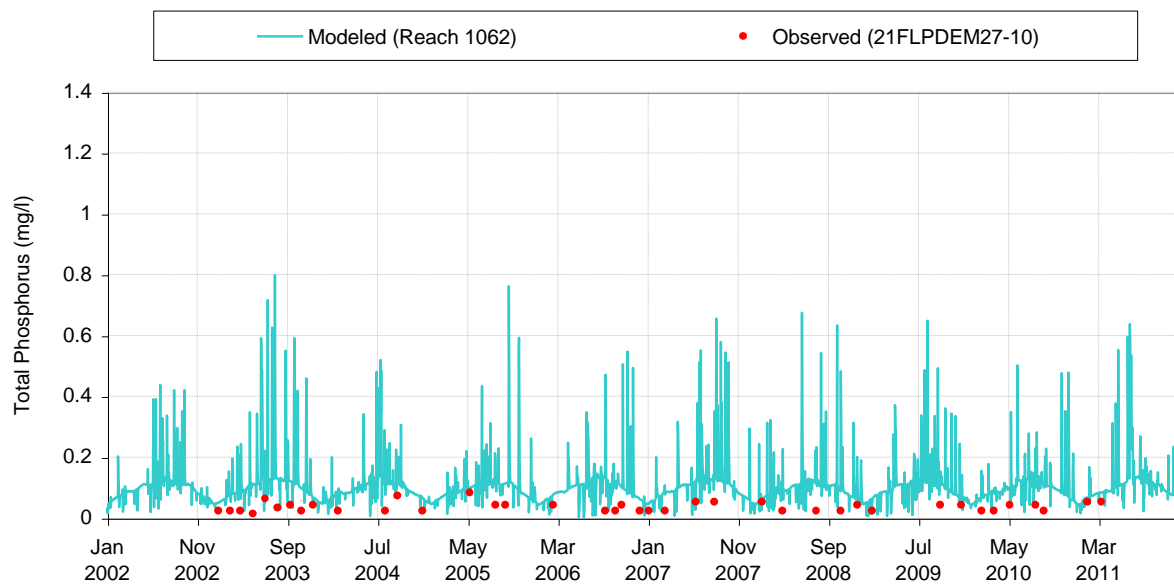


Figure 7.8 Modeled vs. Observed Total Phosphorus (mg/l) at 21FLPDEM27-10

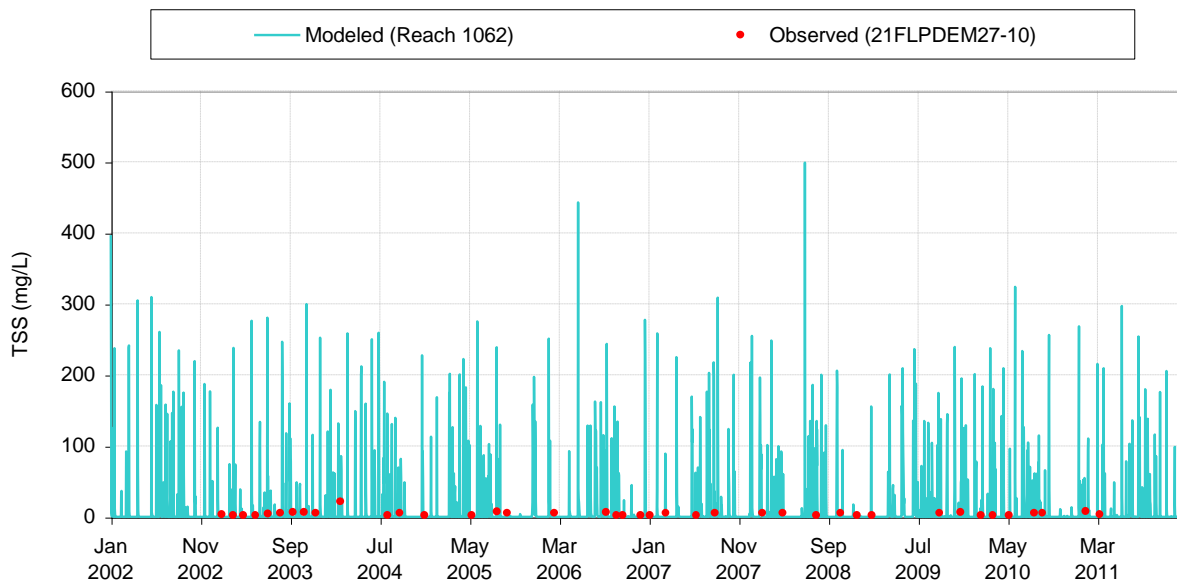


Figure 7.9 Modeled vs. Observed TSS (mg/L) at 21FLPDEM27-10

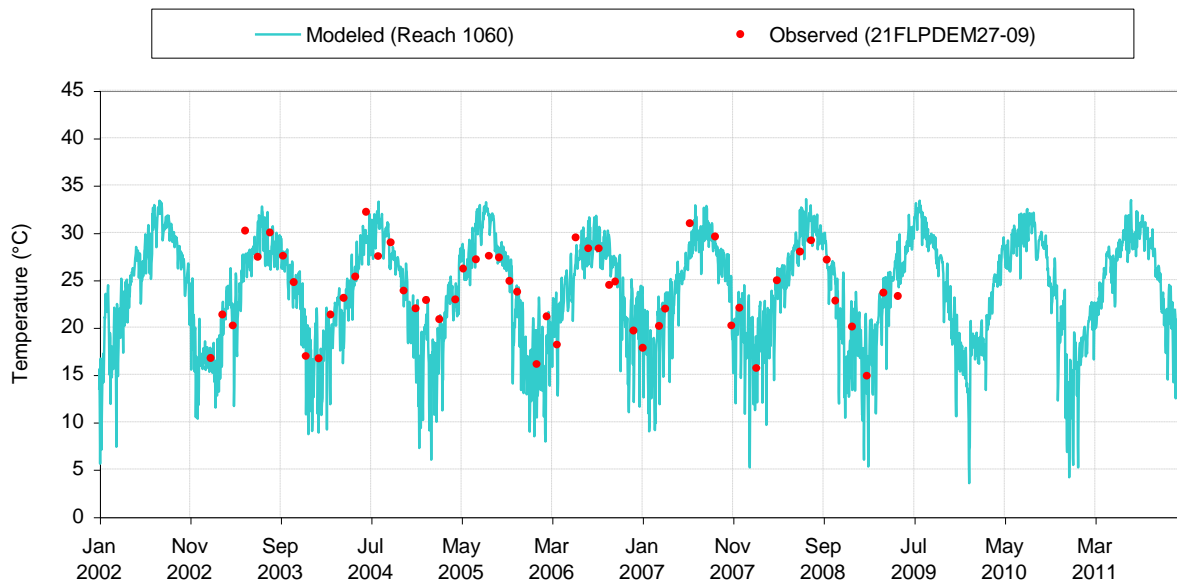


Figure 7.10 Modeled vs. Observed Temperature (°C) at 21FLPDEM27-09

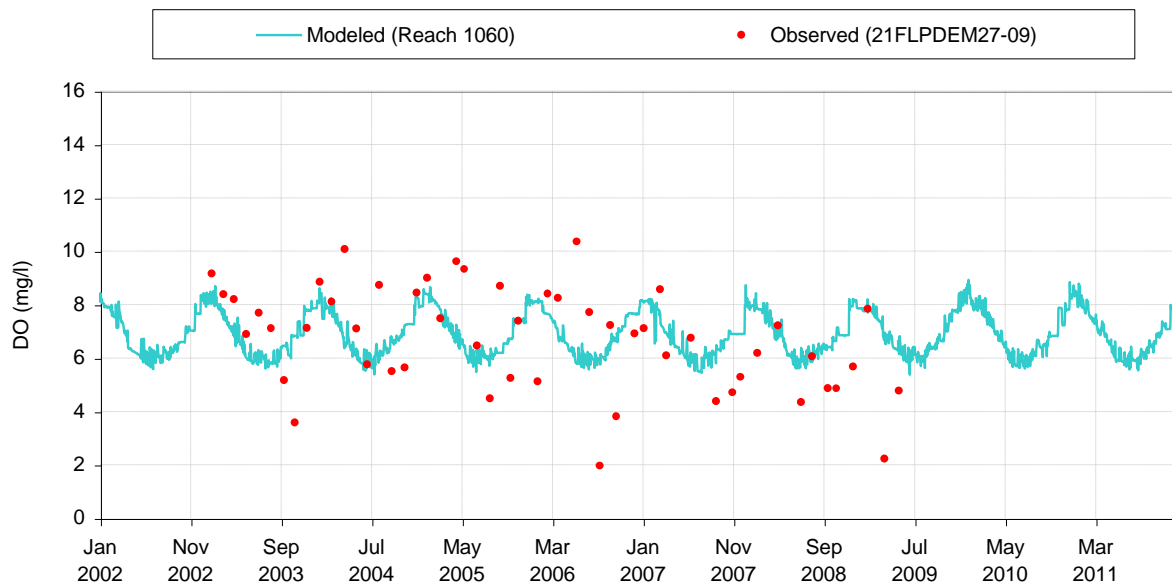


Figure 7.11 Modeled vs. Observed DO (mg/l) at 21FLPDEM27-09

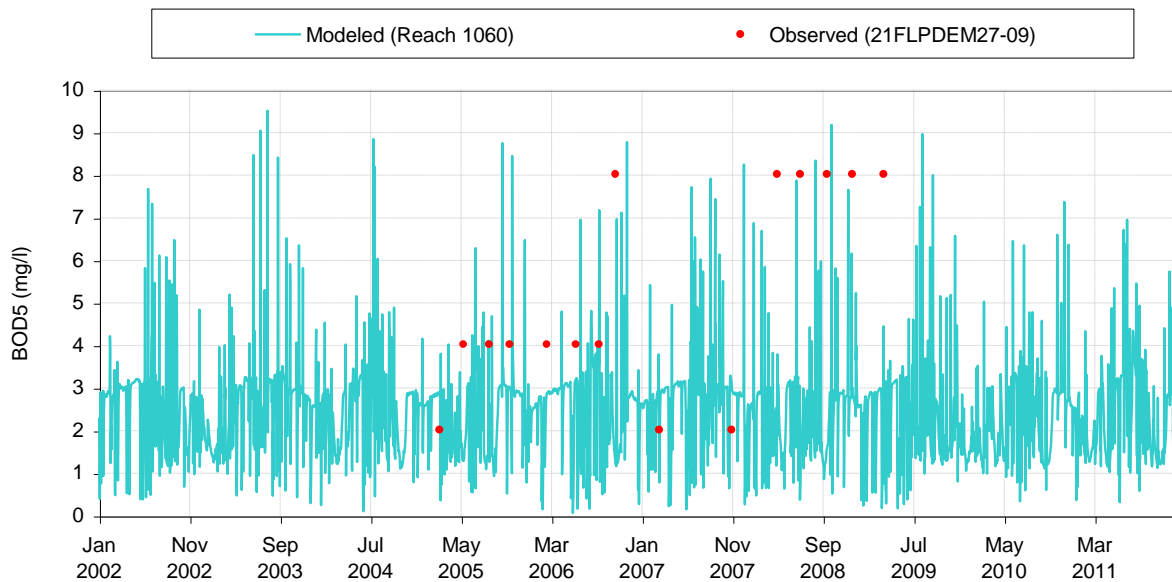


Figure 7.12 Modeled vs. Observed BOD<sub>5</sub> (mg/l) at 21FLPDEM27-09

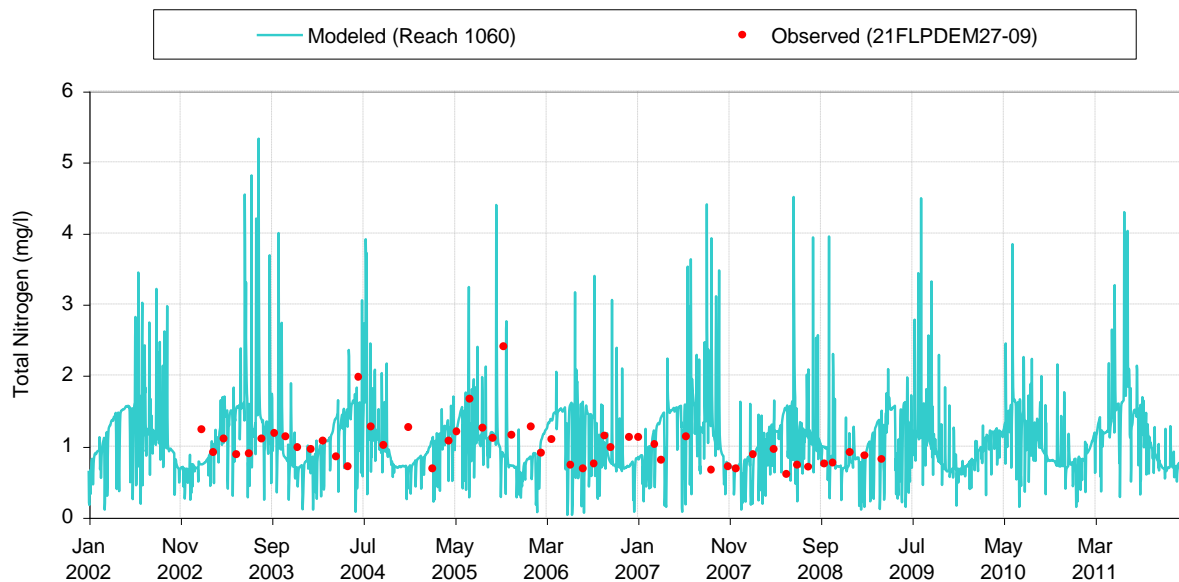


Figure 7.13 Modeled vs. Observed Total Nitrogen (mg/l) at 21FLPDEM27-09

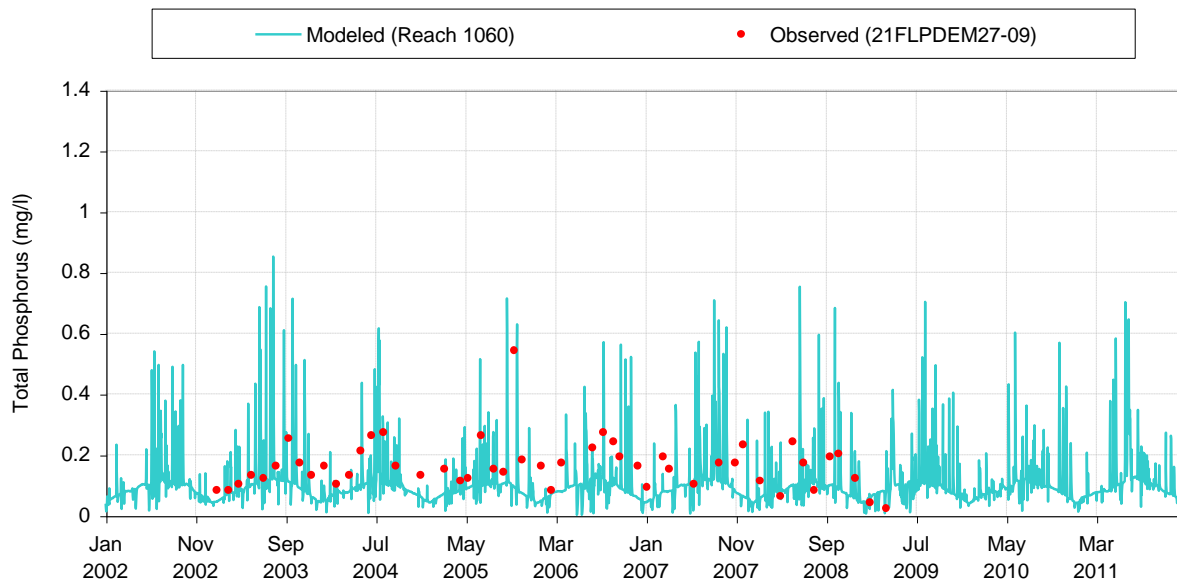


Figure 7.14 Modeled vs. Observed Total Phosphorus (mg/l) at 21FLPDEM27-09

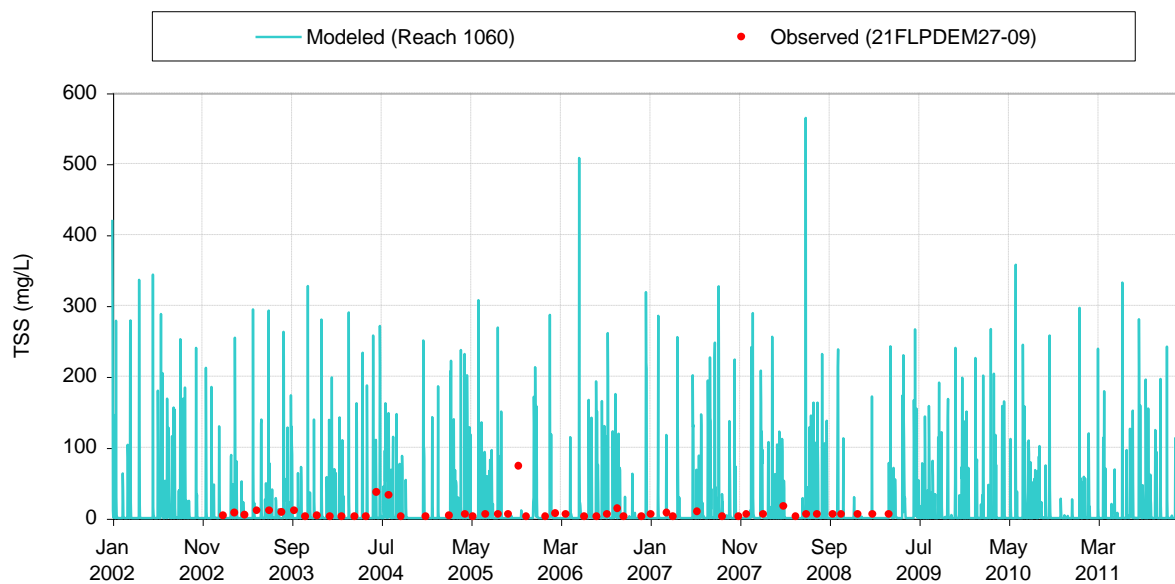


Figure 7.15 Modeled vs. Observed TSS (mg/L) at 21FLPDEM27-09

### 7.1.2 Environmental Fluids Dynamic Code (EFDC)

The EFDC model is a part of the USEPA TMDL Modeling Toolbox due to its application in many TMDL-type projects. As such, the code has been peer reviewed and tested and has been freely distributed and supported by Tetra Tech. EFDC was developed by Dr. John Hamrick (Hamrick 1992) and is currently supported by Tetra Tech for USEPA Office of Research and Development (ORD), USEPA Region 4, and USEPA Headquarters. The models, tools, and databases in the TMDL Modeling Toolbox are continually updated and upgraded through TMDL development in Region 4. EFDC is a multifunctional, surface-water modeling system, which includes hydrodynamic, sediment contaminant, and eutrophication components. The EFDC model is capable of 1, 2, and 3-dimensional spatial resolution. The model employs a curvilinear-orthogonal horizontal grid and a sigma or terrain following vertical grid.

The EFDC hydrodynamic model can run independently of a water quality model. The EFDC model simulates the hydrodynamic and constituent transport and then writes a hydrodynamic linkage file for a water quality model such as the Water Quality Analysis Program (WASP7) model. This model linkage, from EFDC hydrodynamics to WASP water quality, has been applied on many USEPA Region 4 projects in support of TMDLs and has been well tested (Wool et al. 2003).

The EFDC model was utilized to simulate three-dimensional circulation dynamics of hydrodynamic state variables (water surface elevation, salinity, and temperature) in the McKay Creek estuary. The McKay Creek model utilized the Big Bend EFDC model that was created for the Florida Numeric Nutrient Criteria, which was resized to meet the modeling needs of McKay Creek (USEPA 2012c and USEPA 2012d).

An orthogonal, curvilinear grid system consisting of 3995 horizontal cells and 4 equally spaced vertical layers was developed for the Big Bend EFDC model. The grid was developed using Gulf of Mexico bathymetry data. The large grid was reduced in size and scale for the McKay Creek EFDC models.

The EFDC model predicts water surface elevation, salinity, and temperature, in response to a set of multiple factors: wind speed and direction, freshwater discharge, tidal water level fluctuation, rainfall, surface heat flux, and temperature and salinity associated with boundary fluxes. Hourly measurements of atmospheric pressure, dry and wet bulb atmospheric temperatures, rainfall rate, wind speed and direction, and fractional cloud cover were obtained from data collected at station two WBAN stations, Apalachicola and Clearwater, for 2002 through 2011. Solar short wave radiation was calculated using the CE-Qual-W2 method.

The Big Bend Estuary model used hourly water surface elevation time series data from the National Oceanic and Atmospheric Administration (NOAA) tidal stations to simulate tides at the open boundary. Observed temperature data at water quality stations were used to simulate the temperature at the open boundaries, and average salinity in the Gulf of Mexico was used to simulate salinity. The Big Bend Estuary was calibrated to measured NOAA tidal stations, and the Big Bend model was used to simulate the open boundary conditions in the McKay Creek model. The inland boundary grid cells for all three models received LSPC simulated watershed discharges.

The McKay Creek EFDC grid consisted of 64 cells, specifically 32 cells in the horizontal direction and was two layers in the vertical direction (Figure 7.16). The grid was developed using bathymetry data from the larger Big Bend model and NOAA tidal charts. Bathymetry was unavailable for the inland, tidally influenced streams, and channel slope from the USGS digital elevation model was used to estimate slope within the channel. The McKay Creek grid extended from the Clearwater Harbor into McKay Creek and Church Creek.

Because there were no NOAA tidal stations located within the McKay Creek estuary, water surface elevation within the modeled cells could not be directly calibrated. Salinity measurements from IWR44 data were used to review the McKay Creek estuary EFDC calibration. Following model review, the salinity and temperature parameters were adjusted accordingly (Figure 7.17).

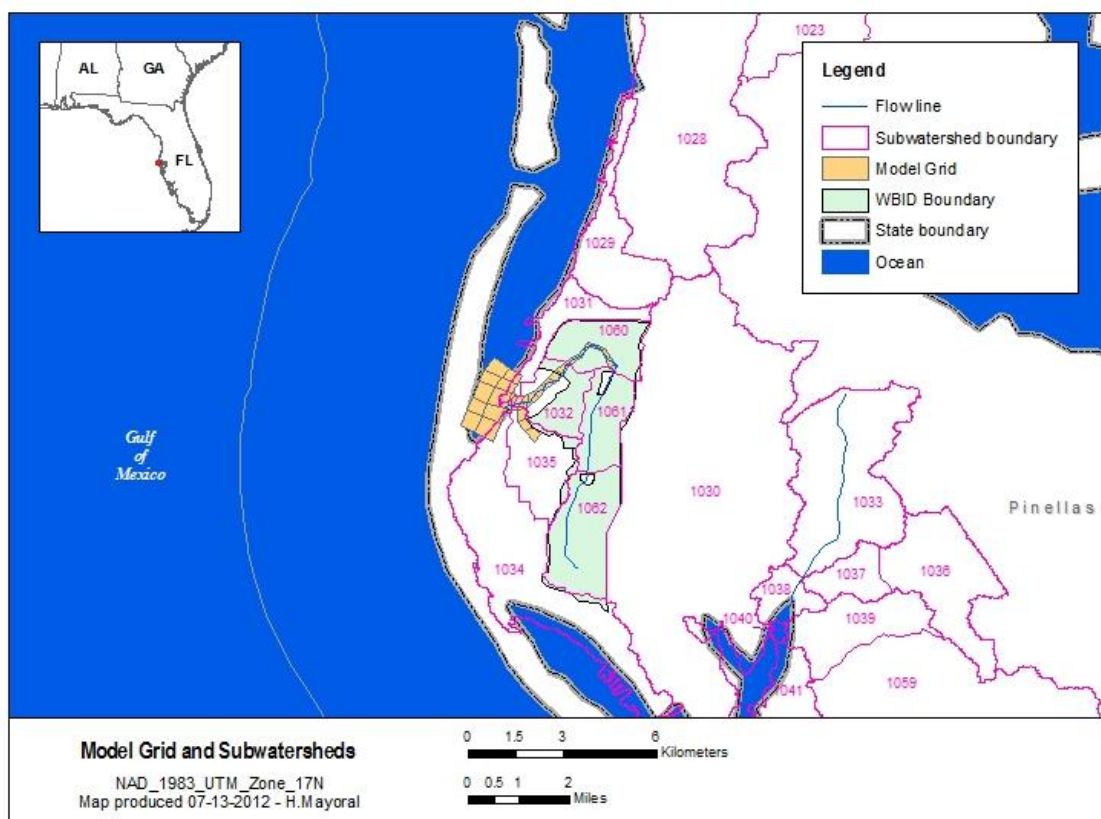


Figure 7.16 Location of McKay Creek LSPC subwatersheds and EFDC grid

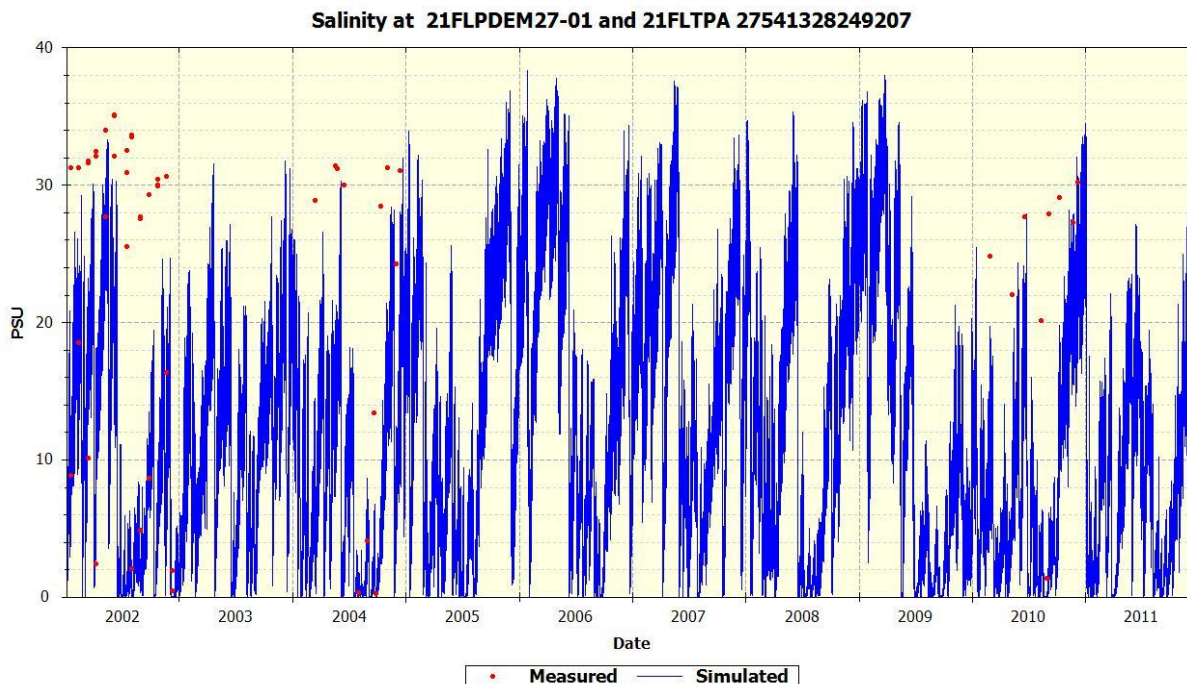


Figure 7.17 Measured versus modeled salinity (PSU) in McKay Creek at station 21FLPDEM27-01 and 21FLTPA 27541328249207

### 7.1.3 Water Quality Analysis Simulation Program (WASP7)

The Water Quality Analysis Simulation Program Version 7.4.1 (WASP7) is an enhanced Windows version of the USEPA Water Quality Analysis Simulation Program (WASP) (Di Toro et al., 1983; Connolly and Winfield, 1984; Ambrose, R.B. et al., 1988), with upgrades to the user's interface and the model's capabilities. The major upgrades to WASP have been the addition of multiple BOD components, addition of sediment diagenesis routines, and addition of periphyton routines. The hydrodynamic file generated by EFDC is compatible with WASP7 and it transfers segment volumes, velocities, temperature and salinity, as well as flows between segments. The time step is set in WASP7 based on the hydrodynamic simulation.

WASP7 helps users interpret and predict water quality responses to natural phenomena and man-made pollution for various pollution management decisions. WASP7 is a dynamic compartment-modeling program for aquatic systems, including both the water column and the underlying benthos. The time-varying processes of advection, dispersion, point and diffuse mass loading and boundary exchange are represented in the basic program. Water quality processes are represented in special kinetic subroutines that are either chosen from a library or written by the user. WASP is structured to permit easy substitution of kinetic subroutines into the overall package to form problem-specific models. WASP7 comes with two such models, TOXI for toxicants and EUTRO for conventional water quality.

The purpose of the WASP7 water quality modeling was to reproduce the three-dimensional transport and chemical and biological interactions of major components of water quality in the McKay Creek estuary. WASP7 modeled total nitrogen (TN) and its speciation, total phosphorus

(TP) and its speciation, chlorophyll-a, dissolved oxygen, and carbonaceous biochemical oxygen demand (CBOD). The model predicts these parameters in response to a set of hydrological, meteorological, atmospheric, and chemical and biological factors: loads from point and nonpoint sources, benthic ammonia and phosphate fluxes, sediment oxygen demand (SOD), solar radiation, air temperature, reaeration, offshore and inland boundary conditions.

The McKay Creek WASP7 model utilized the same grid cells that were developed for the McKay Creek EFDC model. The hydrodynamic simulation from the McKay Creek EFDC model was input into the WASP7 model. Open boundary water quality conditions used measured water quality data from Clearwater Harbor. Water quality loading from the LSPC model was used to simulate loads coming from rivers and streams into the estuary.

Because the LSPC model simulated TN, TP, and BOD and the WASP model simulated TN and its speciation, TP and its speciation, and CBOD, the water quality concentrations from LSPC were adjusted for WASP simulation prior to being input into the WASP model. TN was speciated into nitrate-nitrite (NOX), ammonia (NH<sub>4</sub>), and organic nitrogen (ON), and TP was speciated into orthophosphate (PO<sub>4</sub>) and organic phosphorus (OP). Water quality data in the McKay Creek watershed was reviewed to determine the ratio of NOX, NH<sub>4</sub>, and ON in TN, and the ratio of PO<sub>4</sub> and OP in TP. These ratios were used to develop the partitioning percentages for TN and TP loads from the LSPC model. For the McKay Creek WASP model, 25% of the TN loading was partitioned to NOX, 5% NH<sub>4</sub>, and 70% to ON, while 50% of the TP loading was portioned to both PO<sub>4</sub> and OP. The in-stream BOD loads from LSPC were converted to ultimate CBOD using an f-ratio of 1.5.

Water quality in the Big Bend model was simulated using EFDC and not WASP7. For this reason, water quality parameters from the Tampa Bay WASP7 model were used to populate the McKay Creek WASP7 model. The McKay Creek estuary model calibration was reviewed against water quality data located in IWR44. Following review, the calibration was adjusted accordingly to provide the best existing scenario model calibration for the water quality parameters of concern. Results at select water quality stations are presented in Figure 7.18 through Figure 7.35.

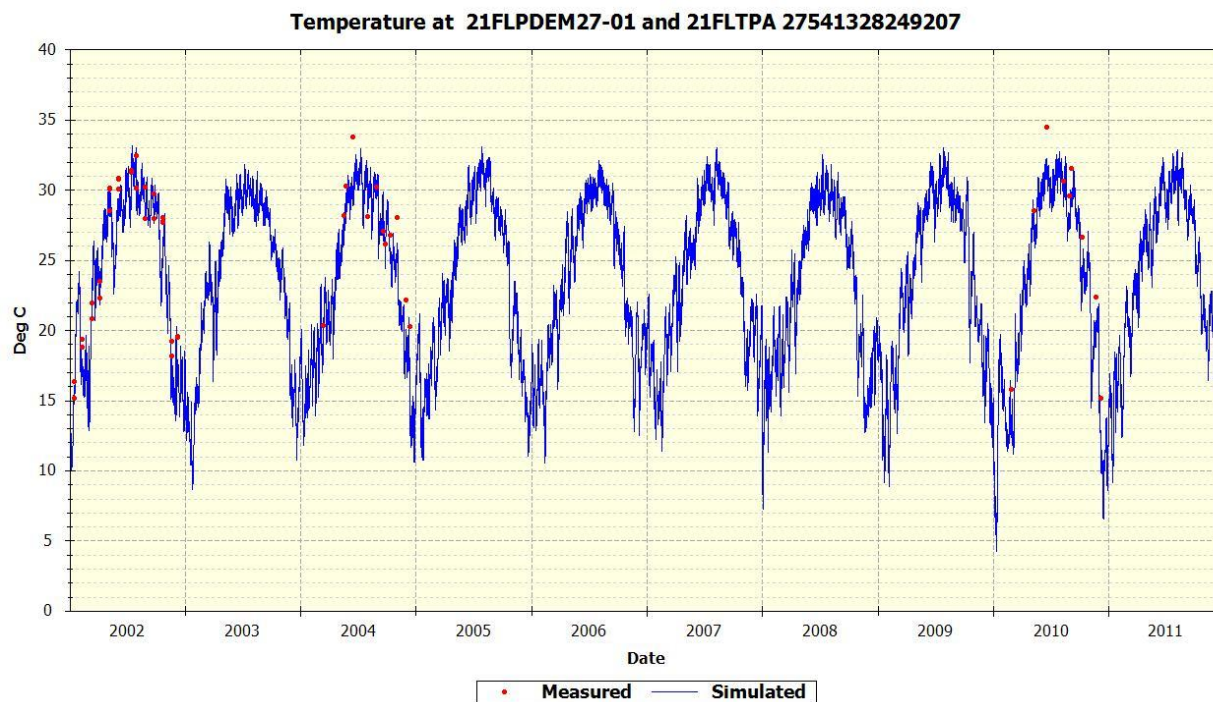


Figure 7.18 Measured verse modeled temperature ( $^{\circ}\text{C}$ ) in McKay Creek at stations 21FLPDEM27-01 and 21FLTPA 27541328249207

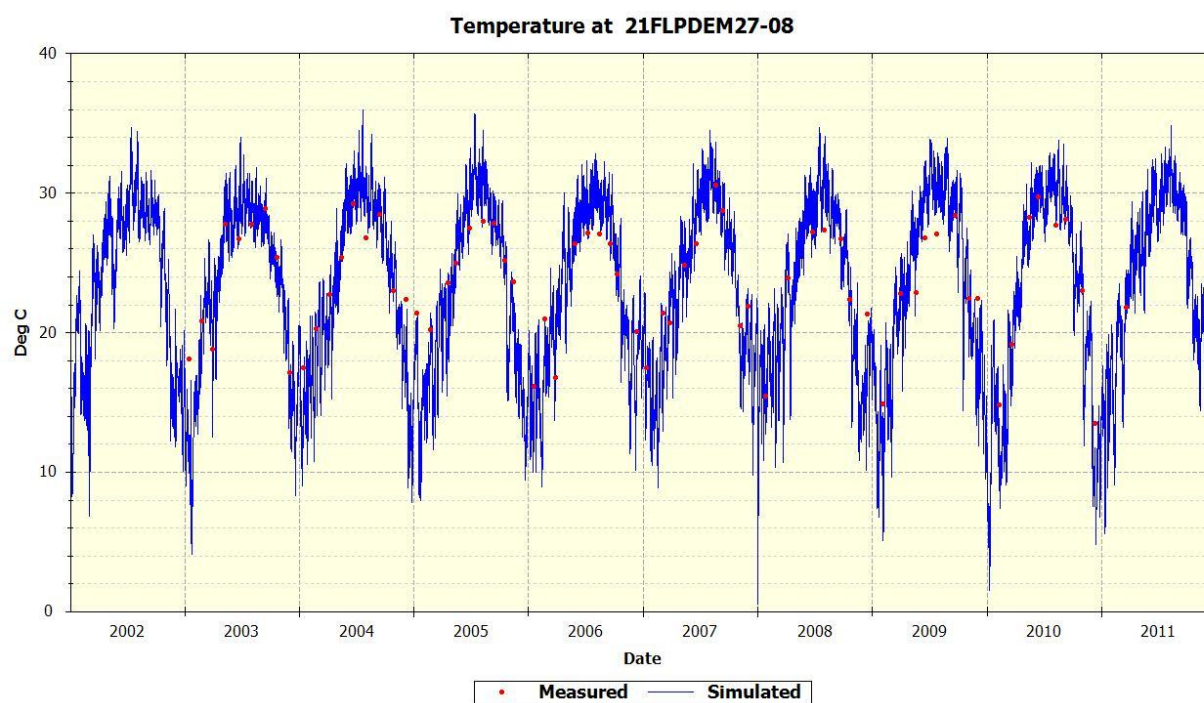


Figure 7.19 Measured verse modeled temperature ( $^{\circ}\text{C}$ ) in Church Creek at station 21FLPDEM27-08

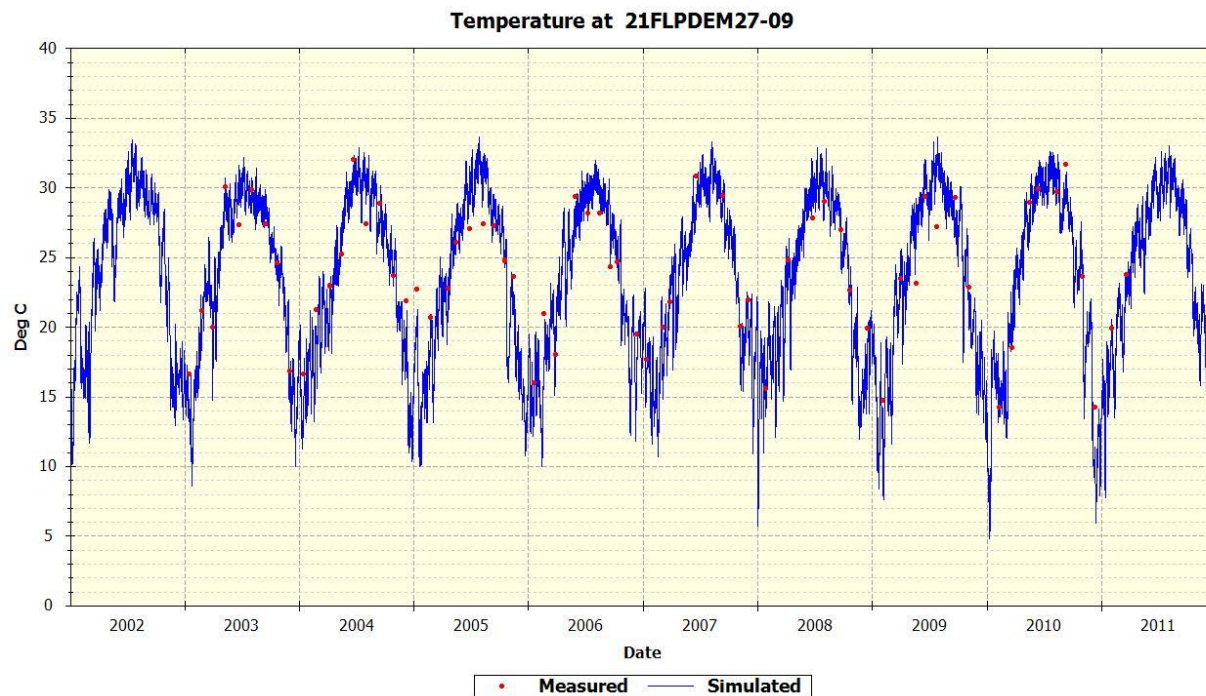


Figure 7.20 Measured versus modeled temperature (°C) in McKay Creek at station 21FLPDEM27-09

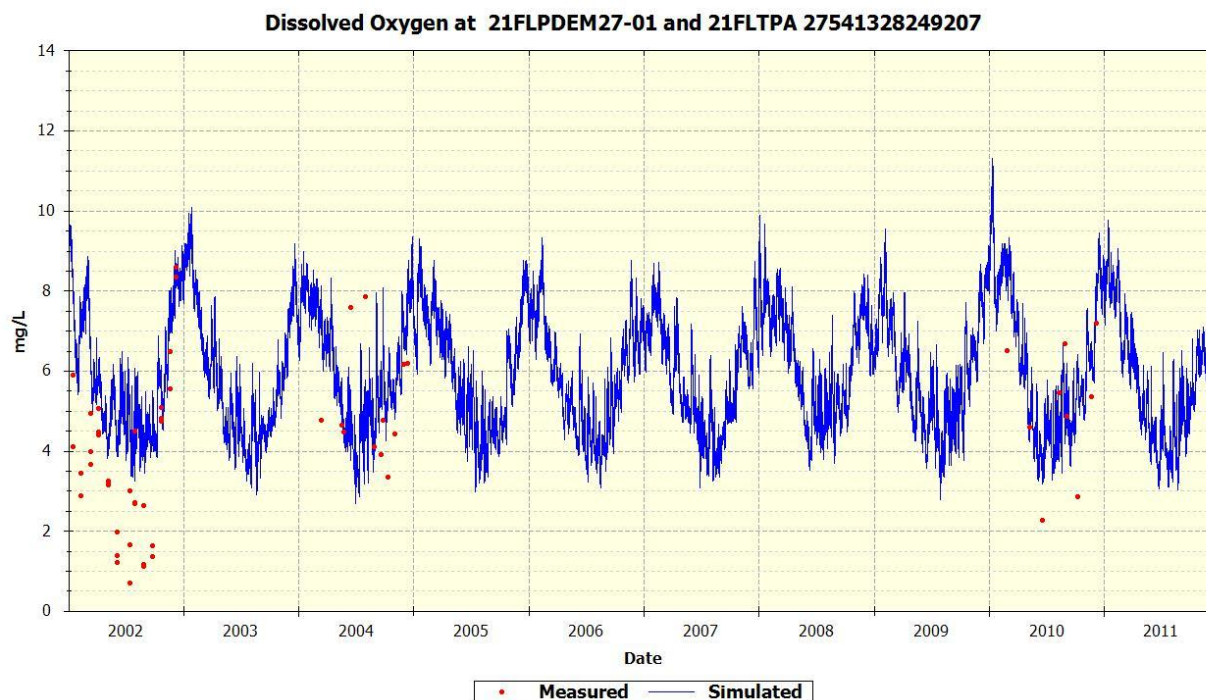


Figure 7.21 Measured versus modeled dissolved oxygen (mg/L) in McKay Creek at stations 21FLPDEM27-01 and 21FLTPA 27541328249207

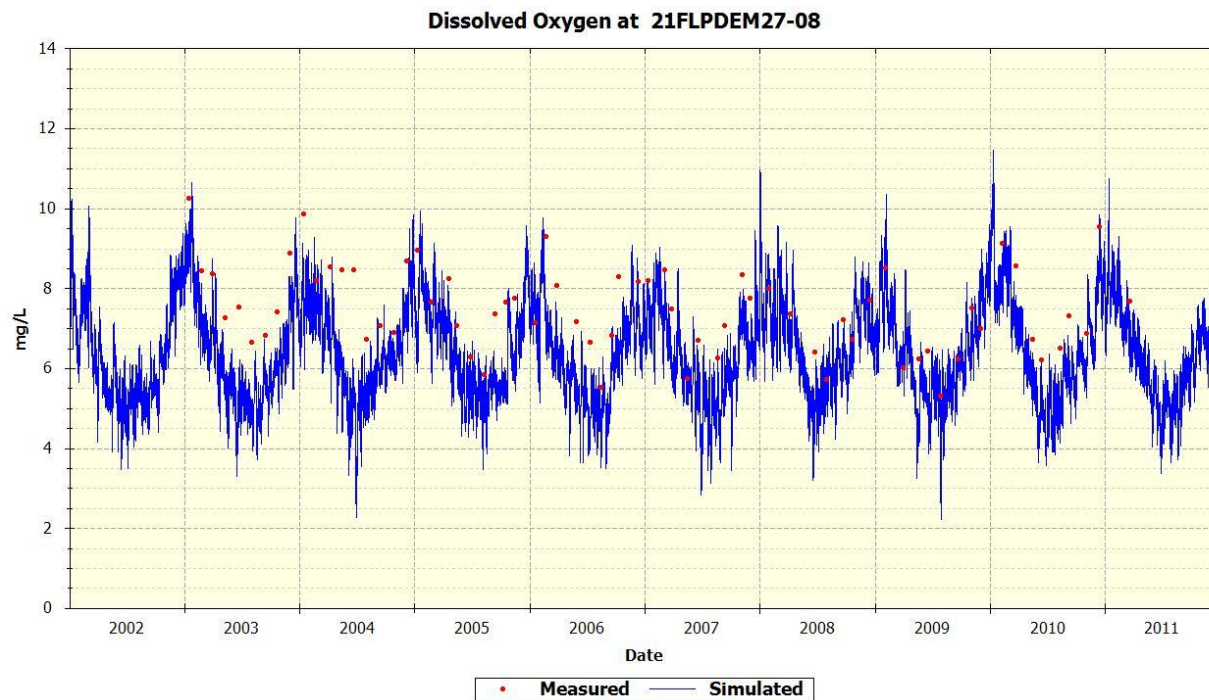


Figure 7.22 Measured versus modeled dissolved oxygen (mg/L) in Church Creek at station 21FLPDEM27-08

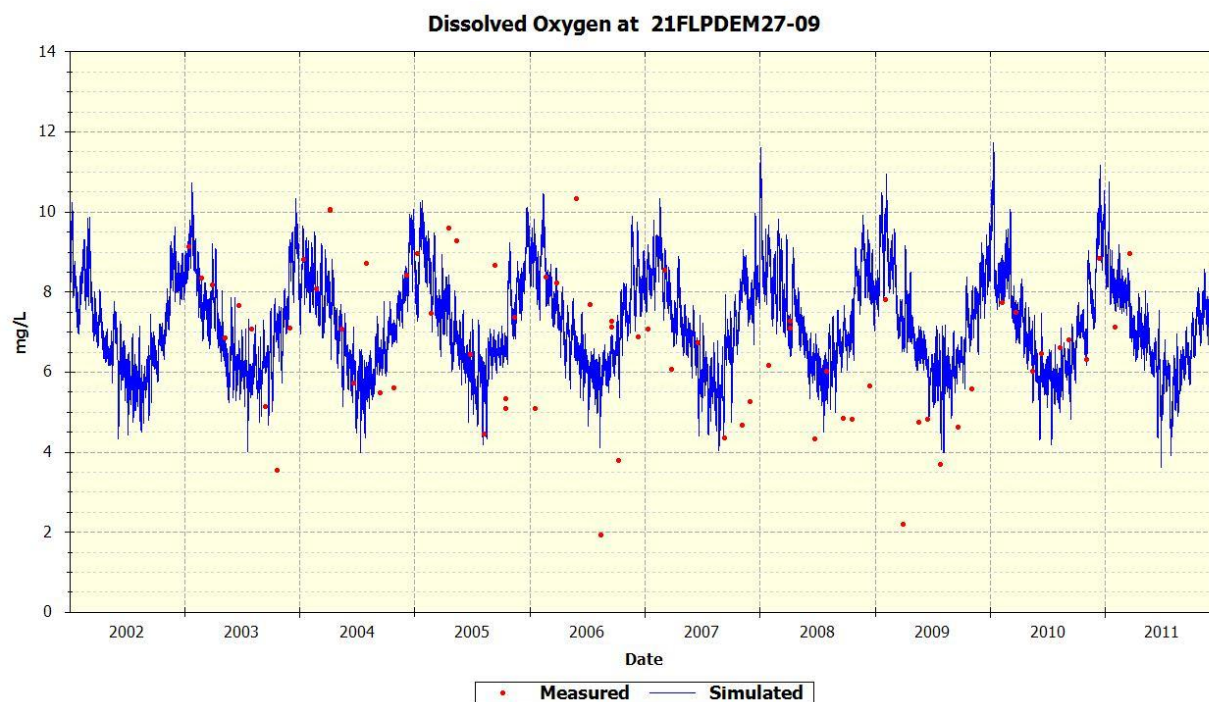


Figure 7.23 Measured versus modeled dissolved oxygen (mg/L) in McKay Creek at station 21FLPDEM27-09

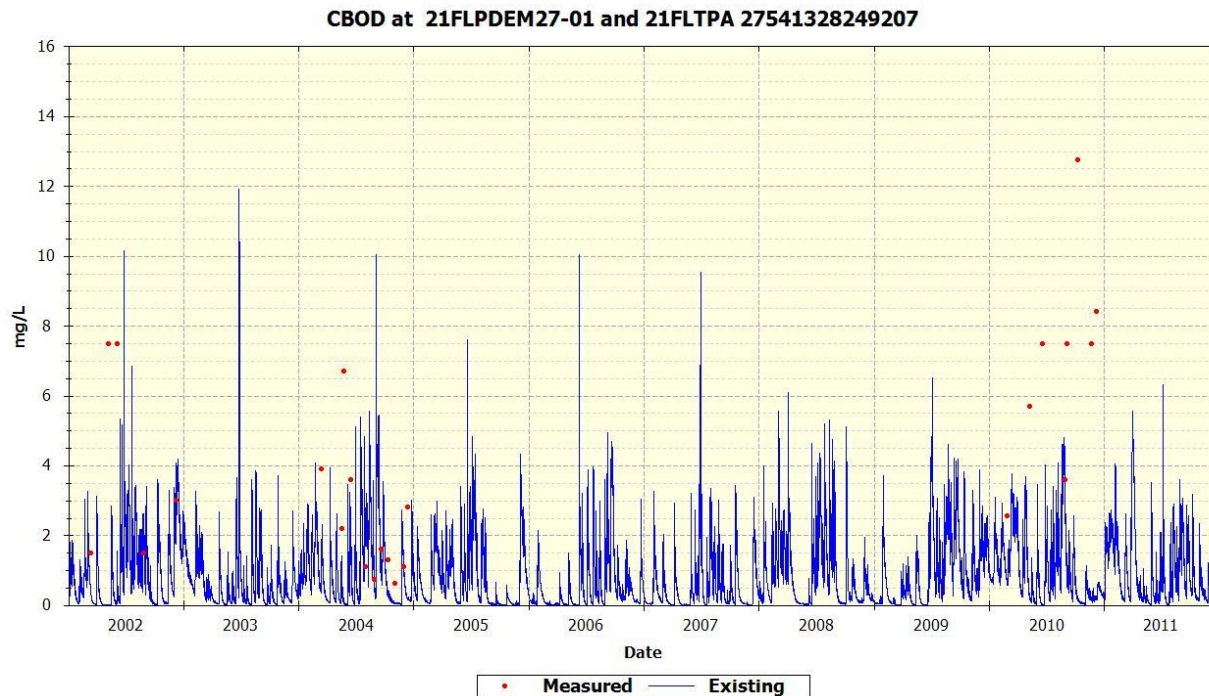


Figure 7.24 Measured verse modeled carbonaceous biochemical oxygen demand (mg/L) in McKay Creek at stations 21FLPDEM27-01 and 21FLTPA 27541328249207

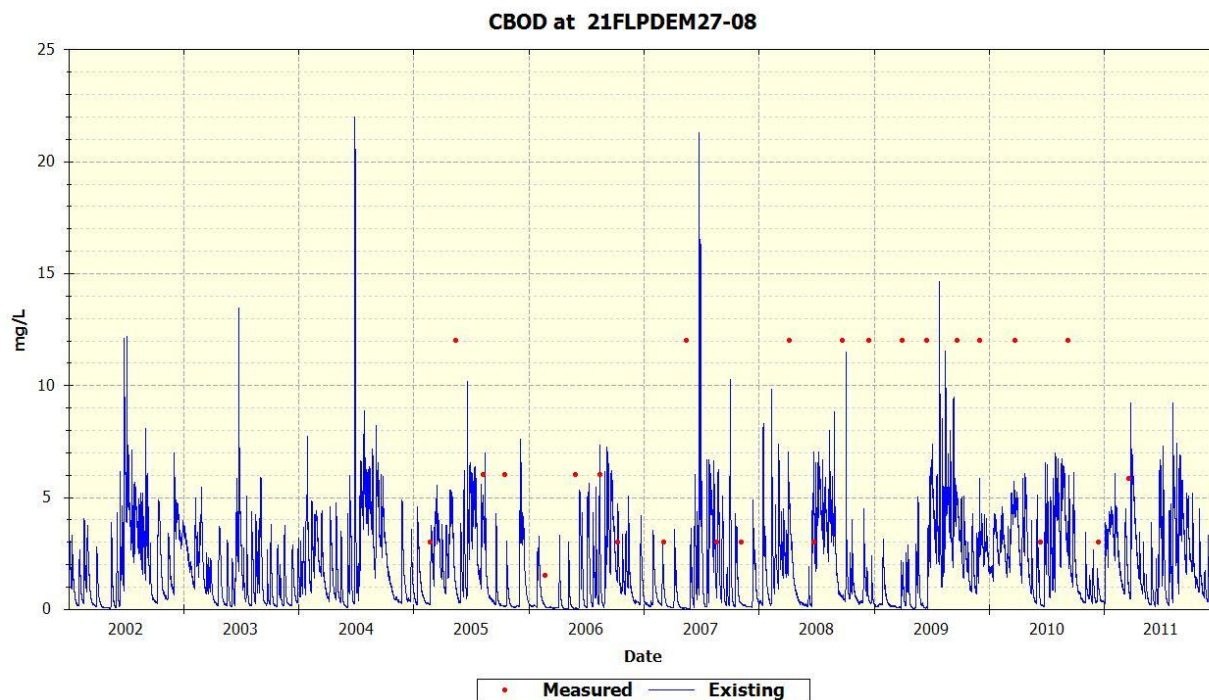


Figure 7.25 Measured verse modeled carbonaceous biochemical oxygen demand (mg/L) in Church Creek at station 21FLPDEM27-08

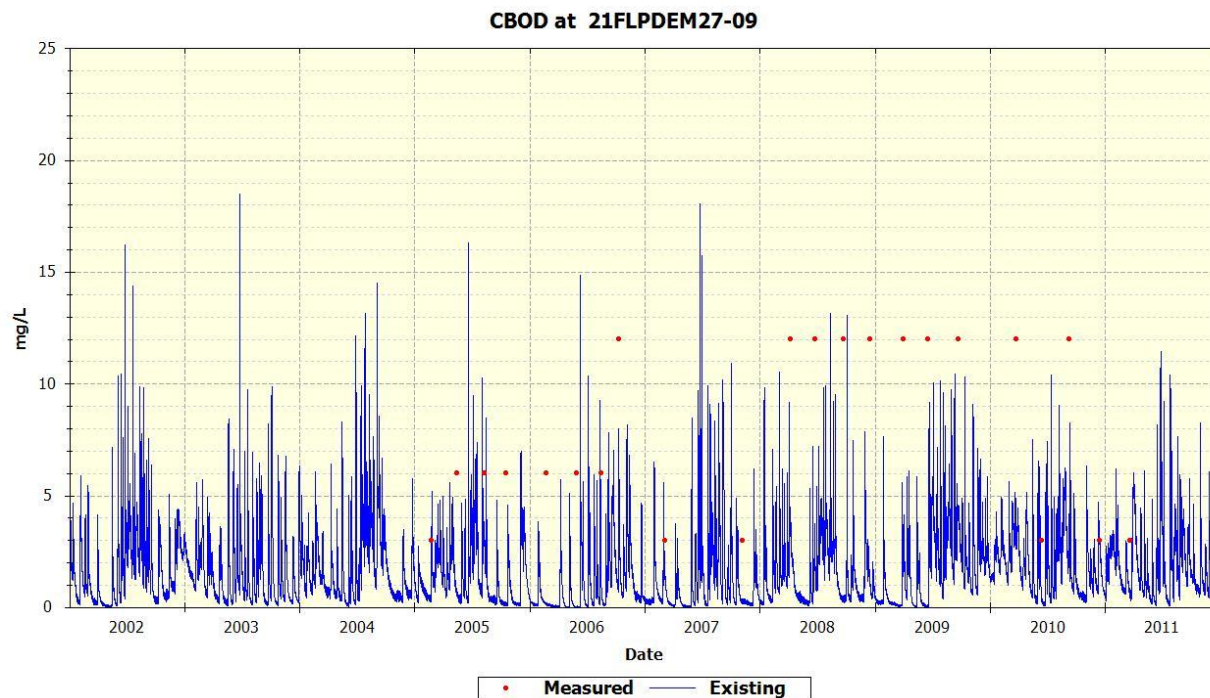


Figure 7.26 Measured verse modeled carbonaceous biochemical oxygen demand (mg/L) in McKay Creek at station 21FLPDEM27-09

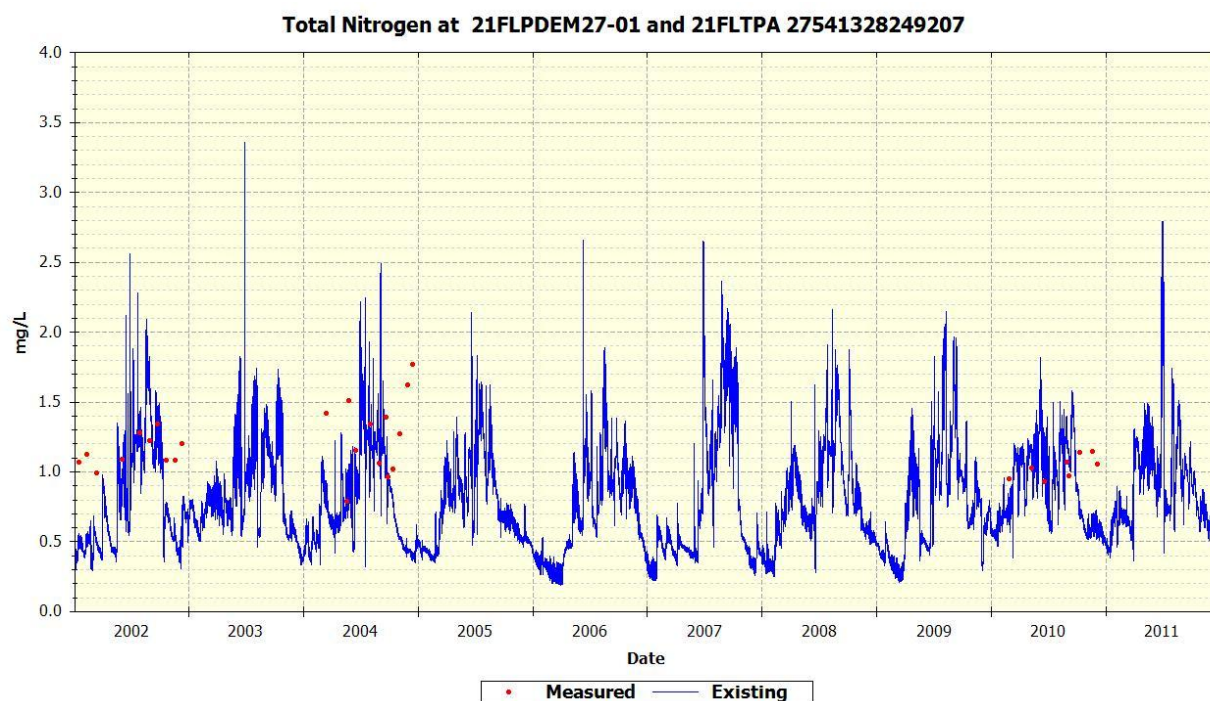


Figure 7.27 Measured verse modeled total nitrogen (mg/L) in McKay Creek at stations 21FLPDEM27-01 and 21FLTPA 27541328249207

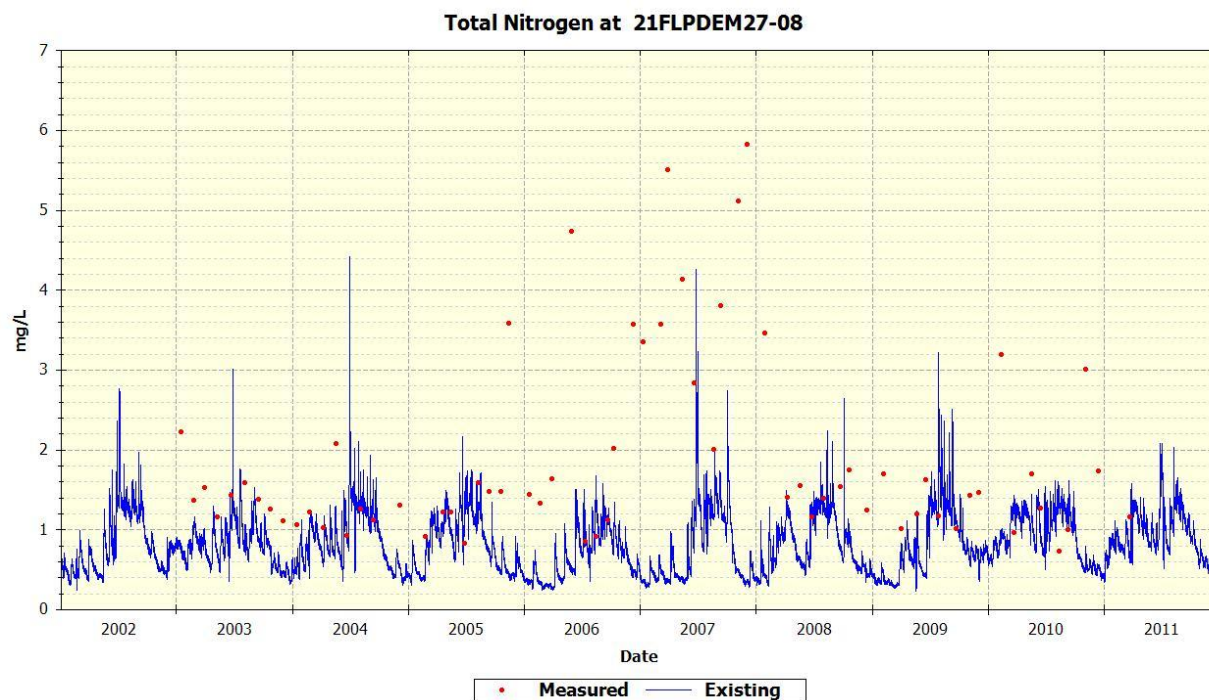


Figure 7.28 Measured verse modeled total nitrogen (mg/L) in Church Creek at station 21FLPDEM27-08

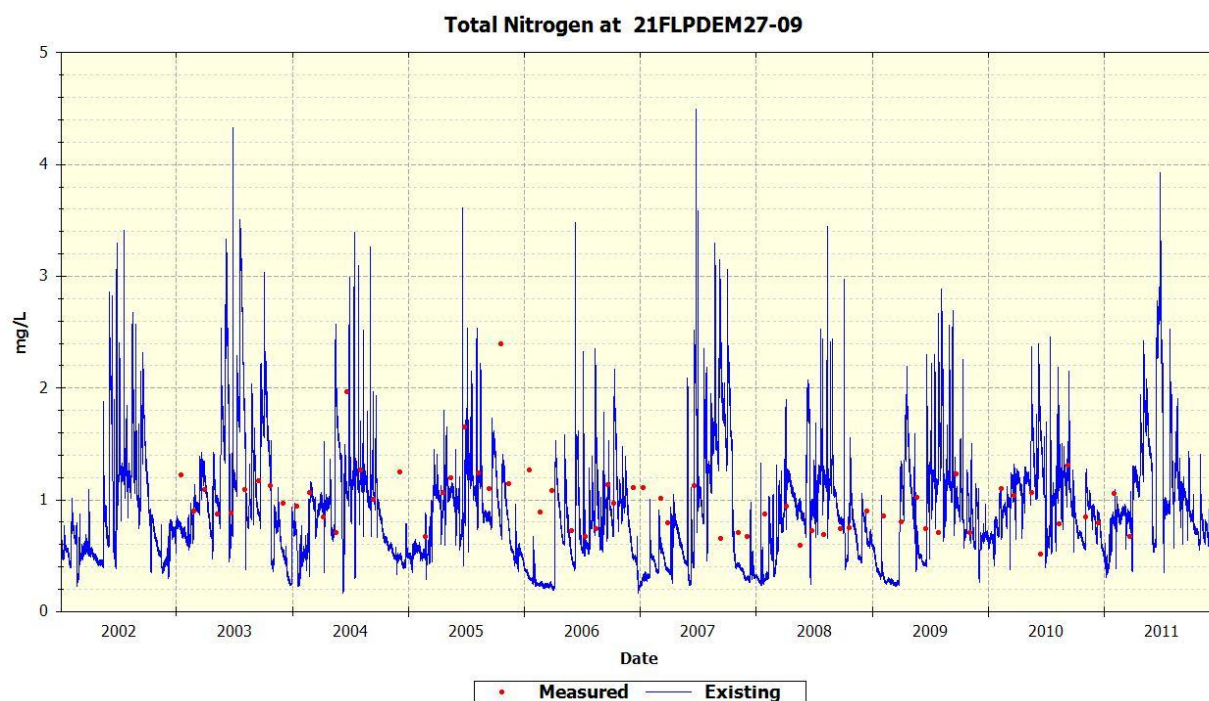


Figure 7.29 Measured verse modeled total nitrogen (mg/L) in McKay Creek at station 21FLPDEM27-09

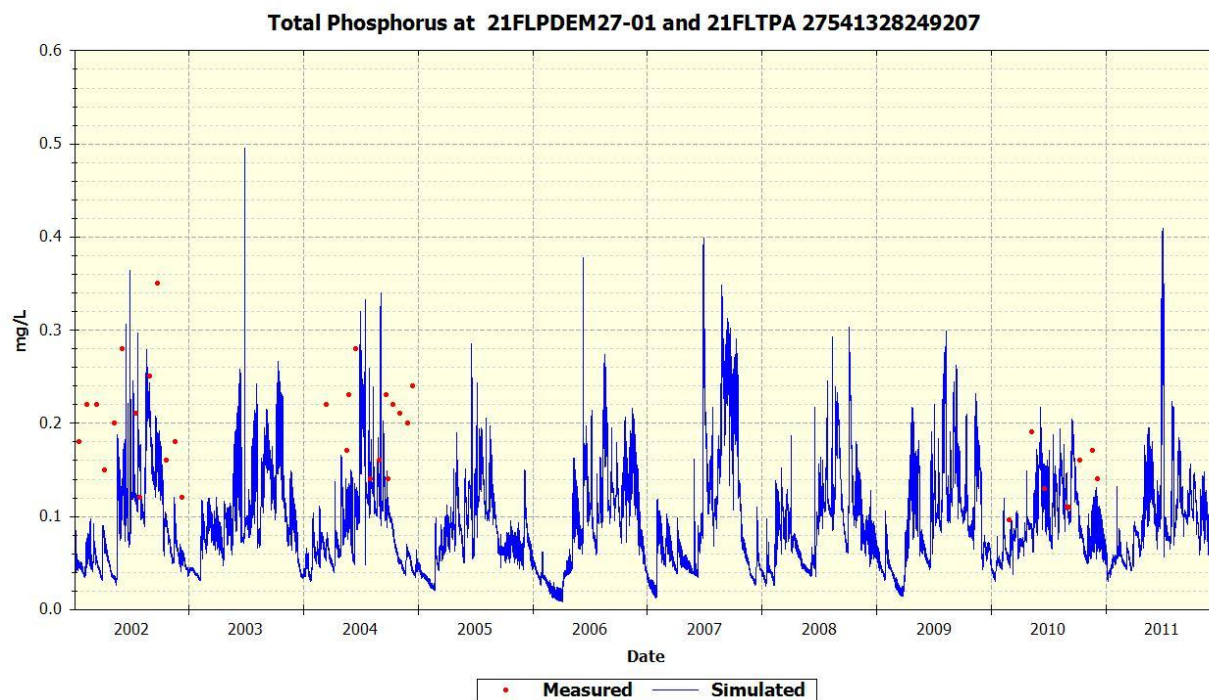


Figure 7.30 Measured verse modeled total phosphorus (mg/L) in McKay Creek at stations 21FLPDEM27-01 and 21FLTPA 27541328249207

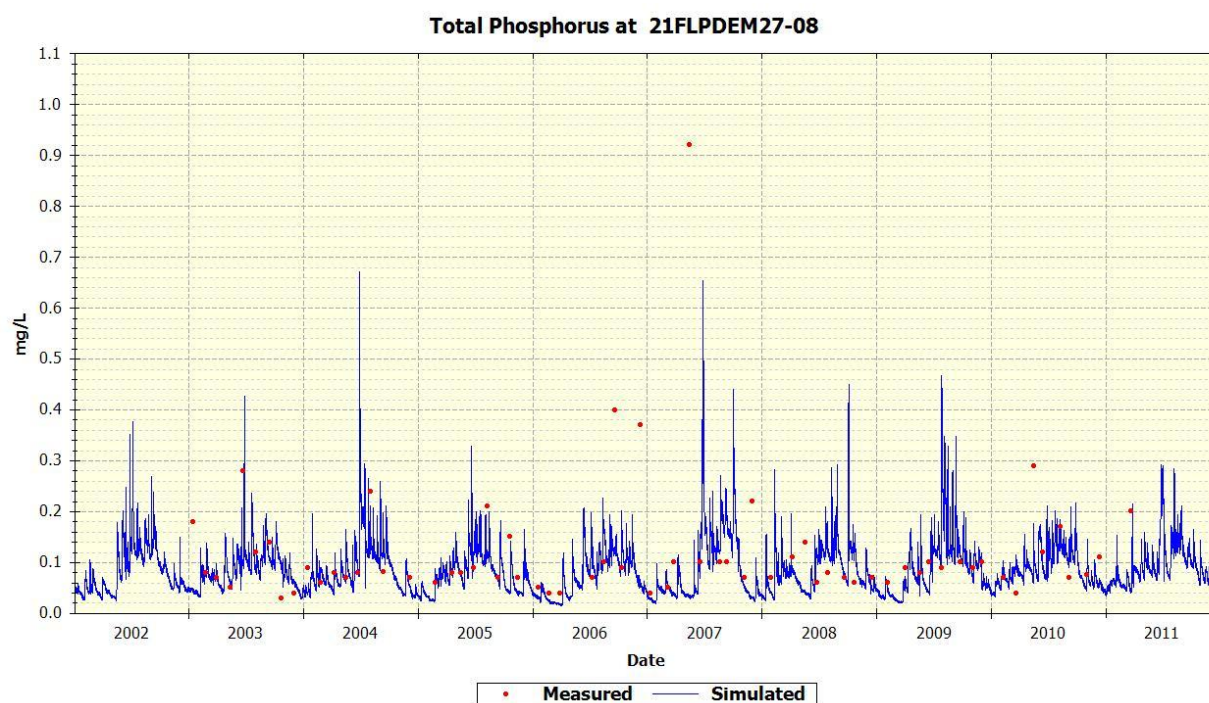


Figure 7.31 Measured verse modeled total phosphorus (mg/L) in Church Creek at station 21FLPDEM27-08

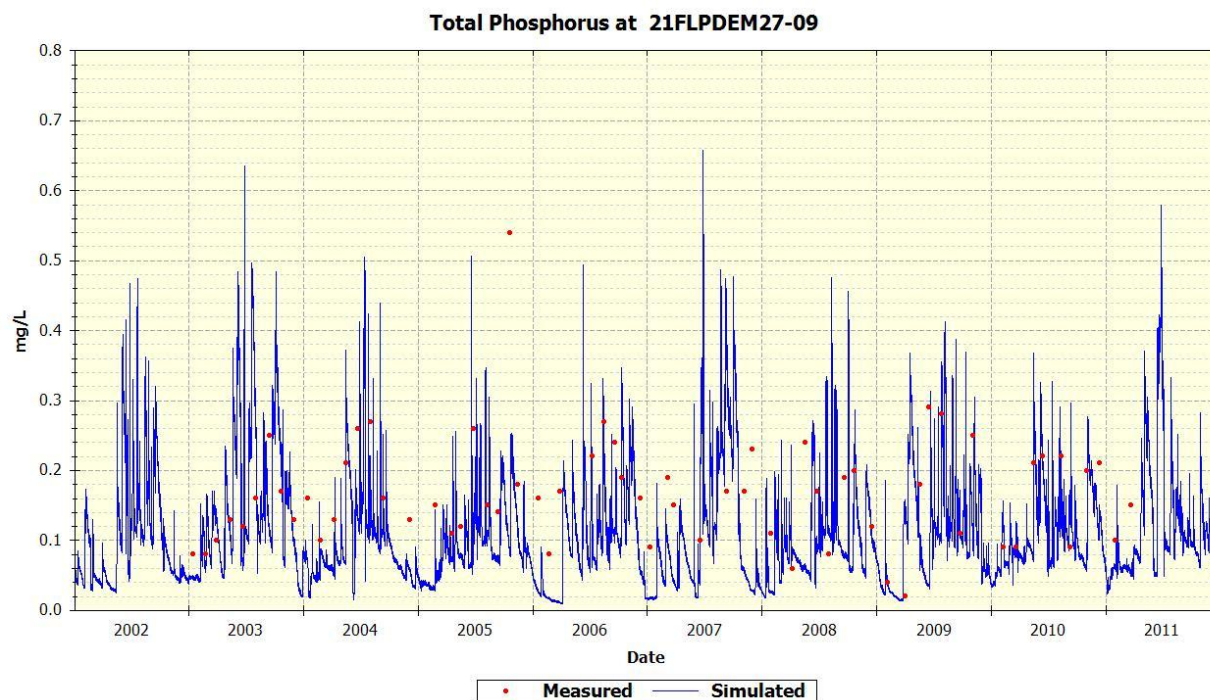


Figure 7.32 Measured verse modeled total phosphorus (mg/L) in McKay Creek at station 21FLPDEM27-09

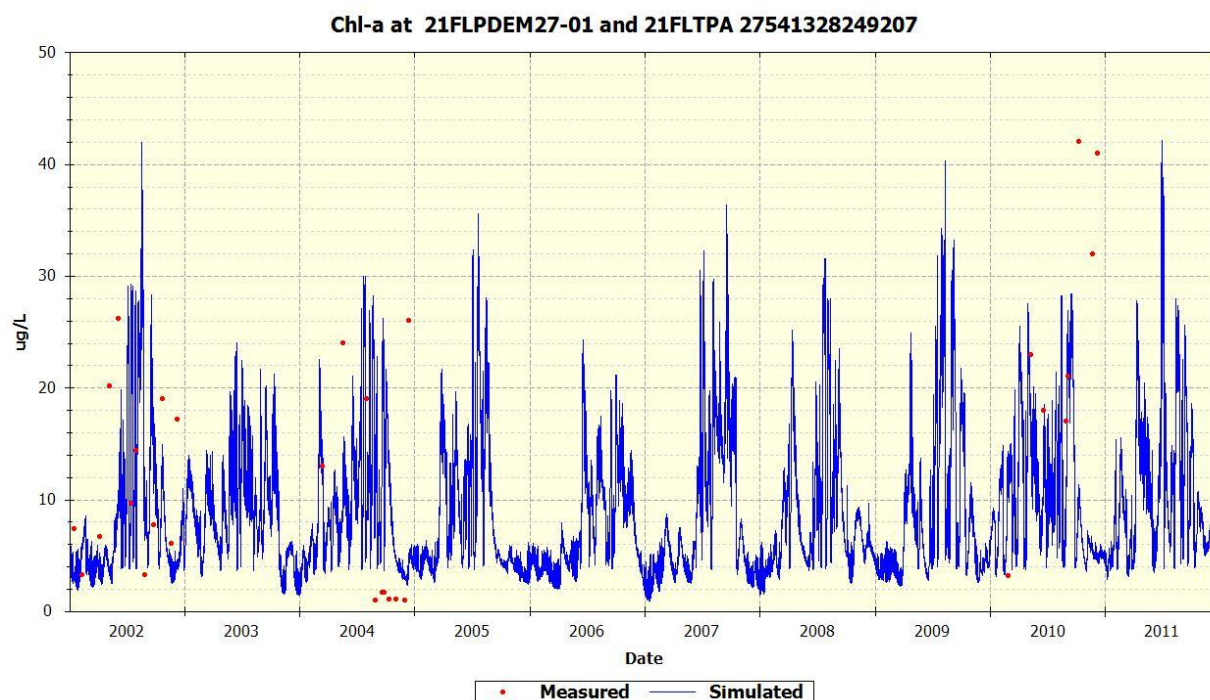


Figure 7.33 Measured verse modeled chlorophyll-a (ug/L) in McKay Creek at stations 21FLPDEM27-01 and 21FLTPA 27541328249207

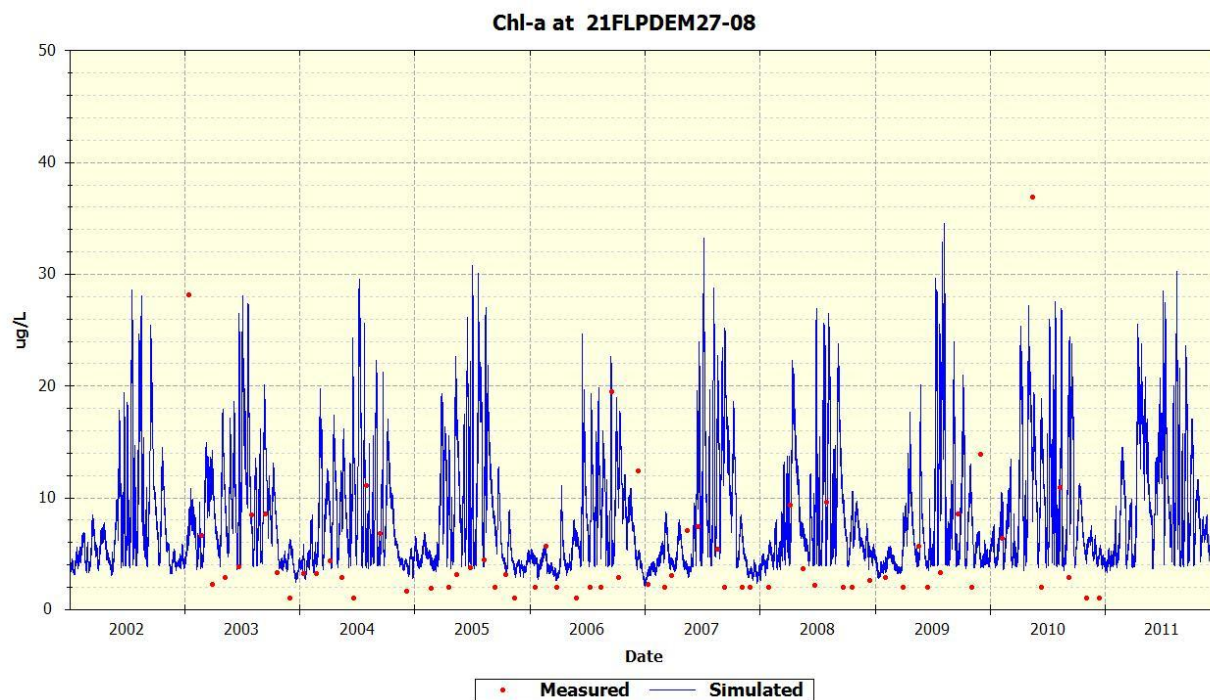


Figure 7.34 Measured versus modeled chlorophyll-a (ug/L) in Church Creek at station 21FLPDEM27-08

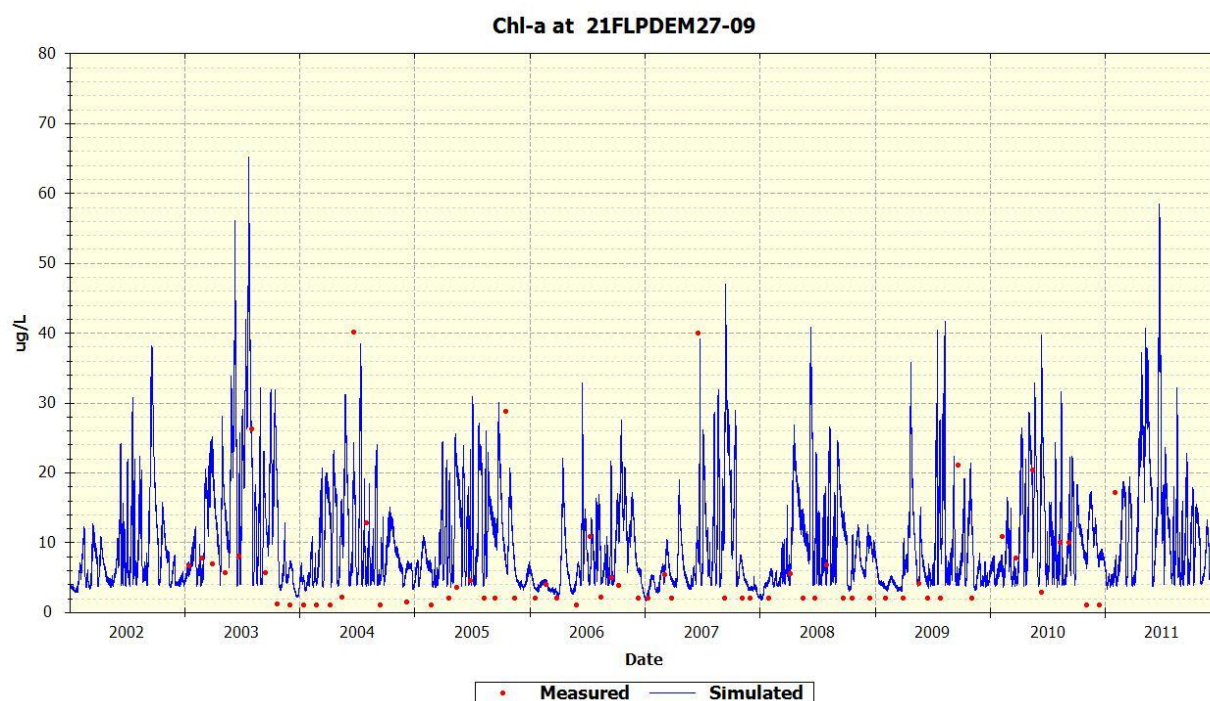


Figure 7.35 Measured versus modeled chlorophyll-a (ug/L) in McKay Creek at station 21FLPDEM27-09

## 7.2 Scenarios

Two modeling scenarios were developed and evaluated in this TMDL determination: a current condition and a natural condition scenario. Concentrations and loadings were evaluated to determine if DO concentrations in the natural condition scenario could meet the DO standard, and the impact of nutrients on the DO concentrations. The results from the scenarios were used to develop the TMDL.

### 7.2.1 Current Condition

The current condition scenario evaluated current hydrologic and water quality conditions in the watershed, specifically water quality concentration and loadings at the outlet of 1633B. The current condition annual average concentrations for the McKay Creek WBID are presented in Table 7.1. The current condition simulation was used to determine the base loadings for the WBID. These base loadings (Table 7.2), when compared with the TMDL scenarios, were used to determine the percent reduction in nutrient loads that will be needed to achieve water quality standards. Figure 7.36 through Figure 7.41 provide the calibrated current condition modeled parameters for WBID 1633B.

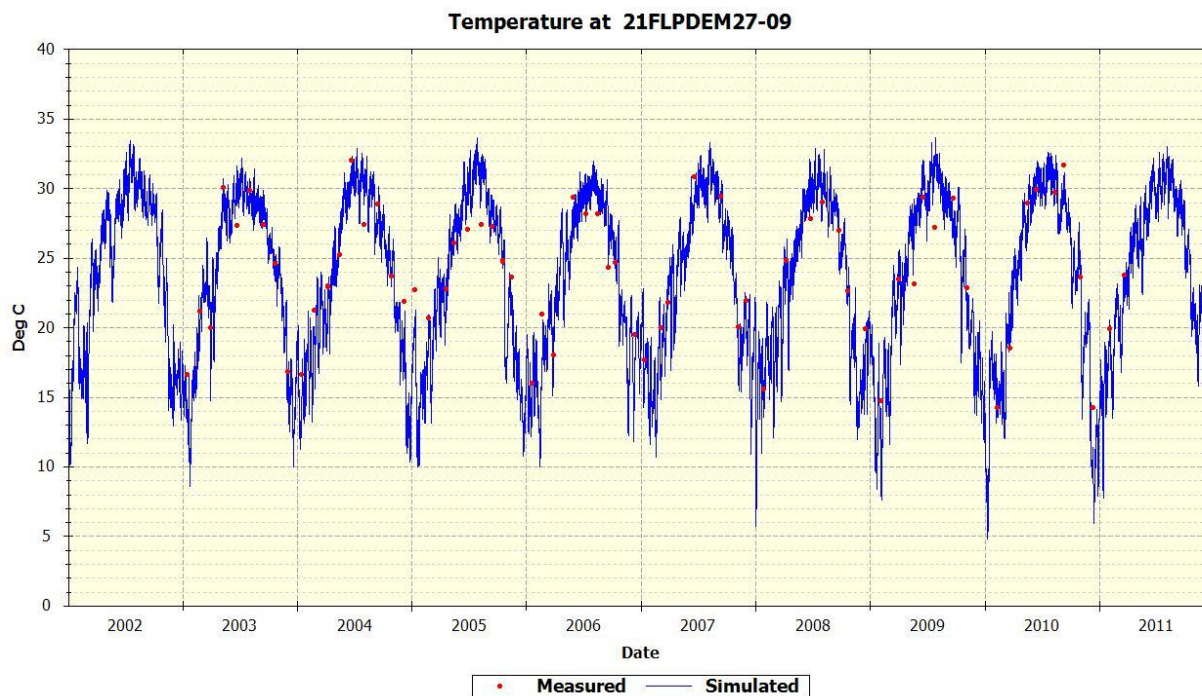


Figure 7.36 Measured verse modeled temperature (°C) in McKay Creek at station 21FLPDEM27-09

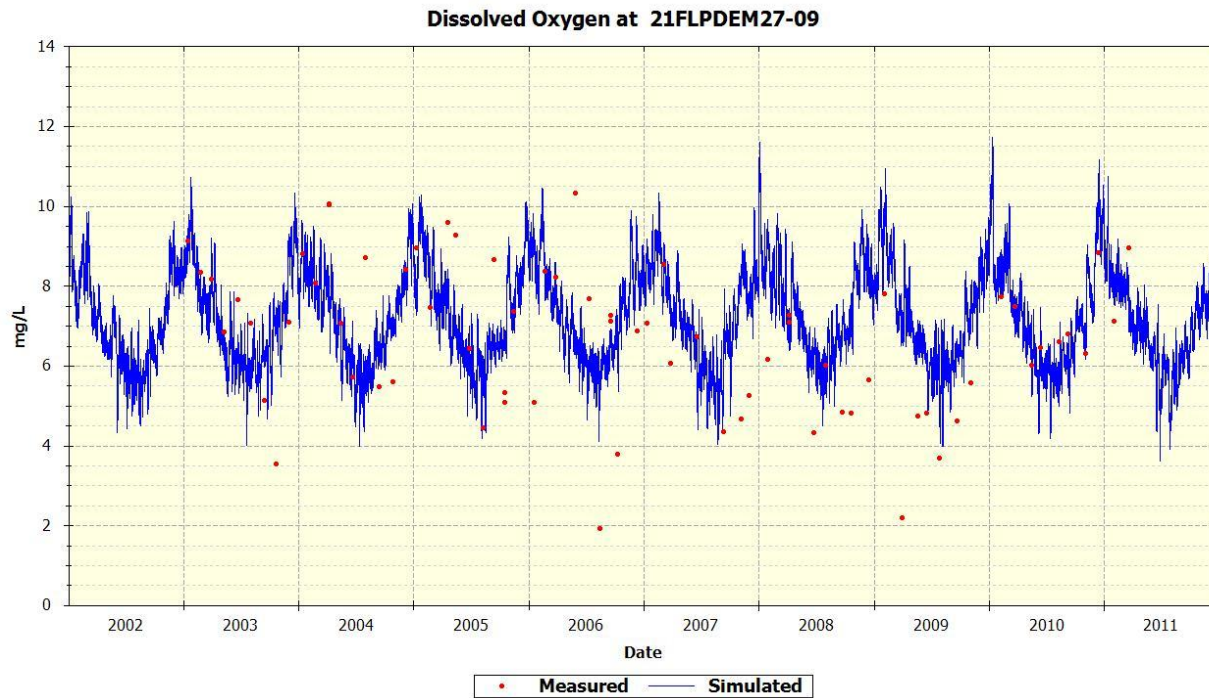


Figure 7.37 Measured verse modeled dissolved oxygen (mg/L) in McKay Creek at station 21FLPDEM27-09

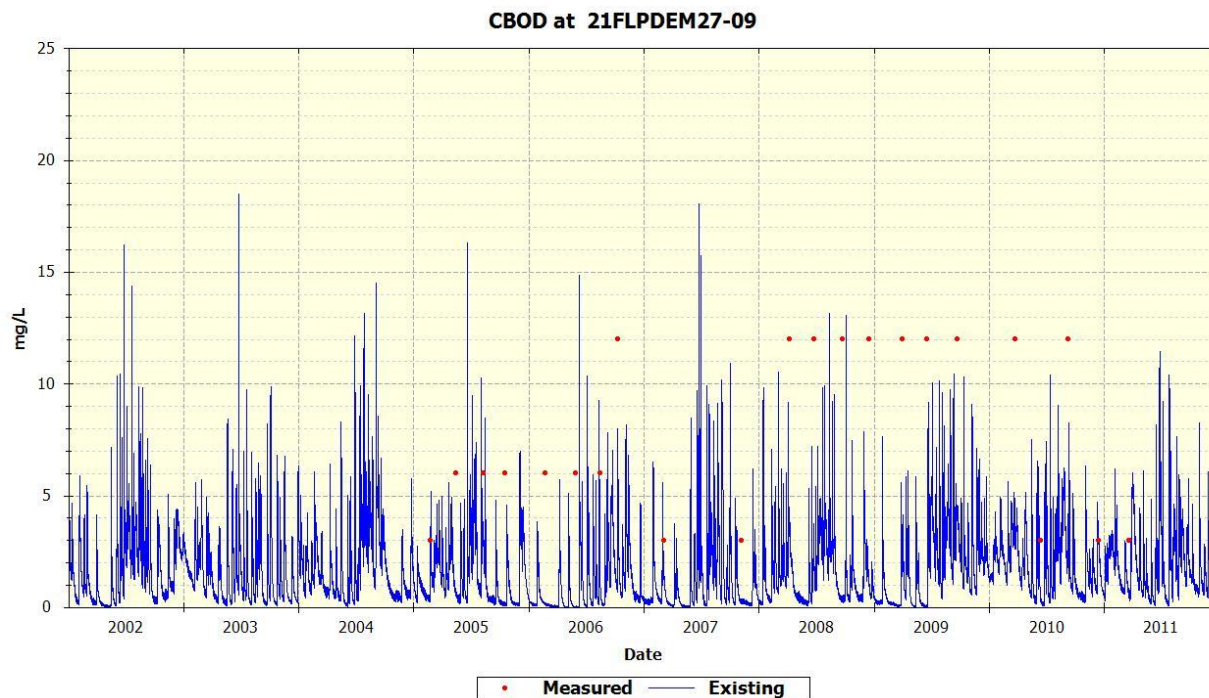


Figure 7.38 Measured verse modeled carbonaceous biochemical oxygen demand (mg/L) in McKay Creek at station 21FLPDEM27-09

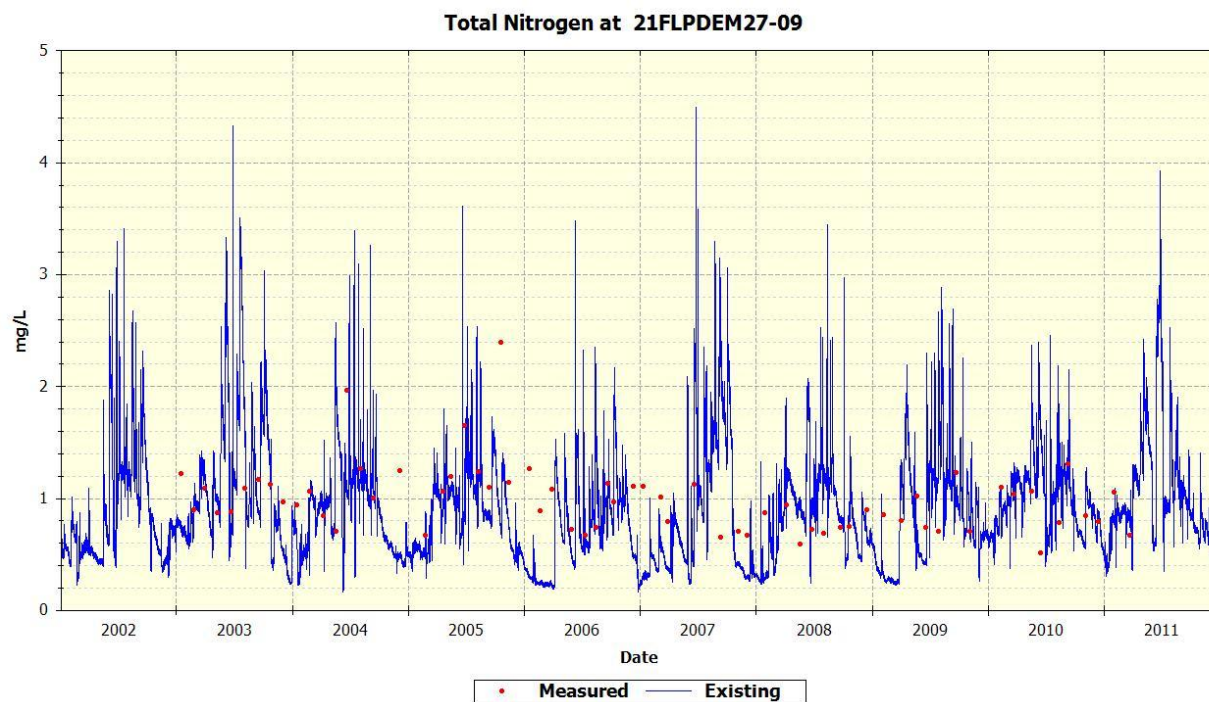


Figure 7.39 Measured verse modeled total nitrogen (mg/L) in McKay Creek at station 21FLPDEM27-09

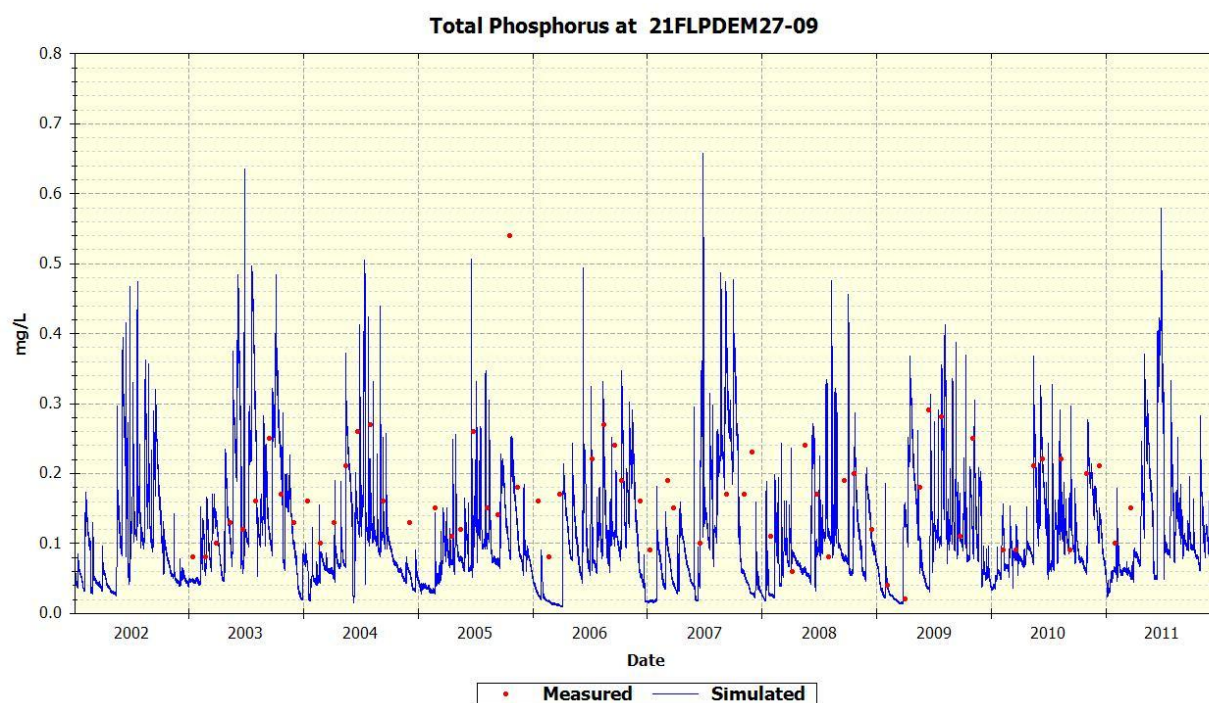


Figure 7.40 Measured verse modeled total phosphorus (mg/L) in McKay Creek at station 21FLPDEM27-09

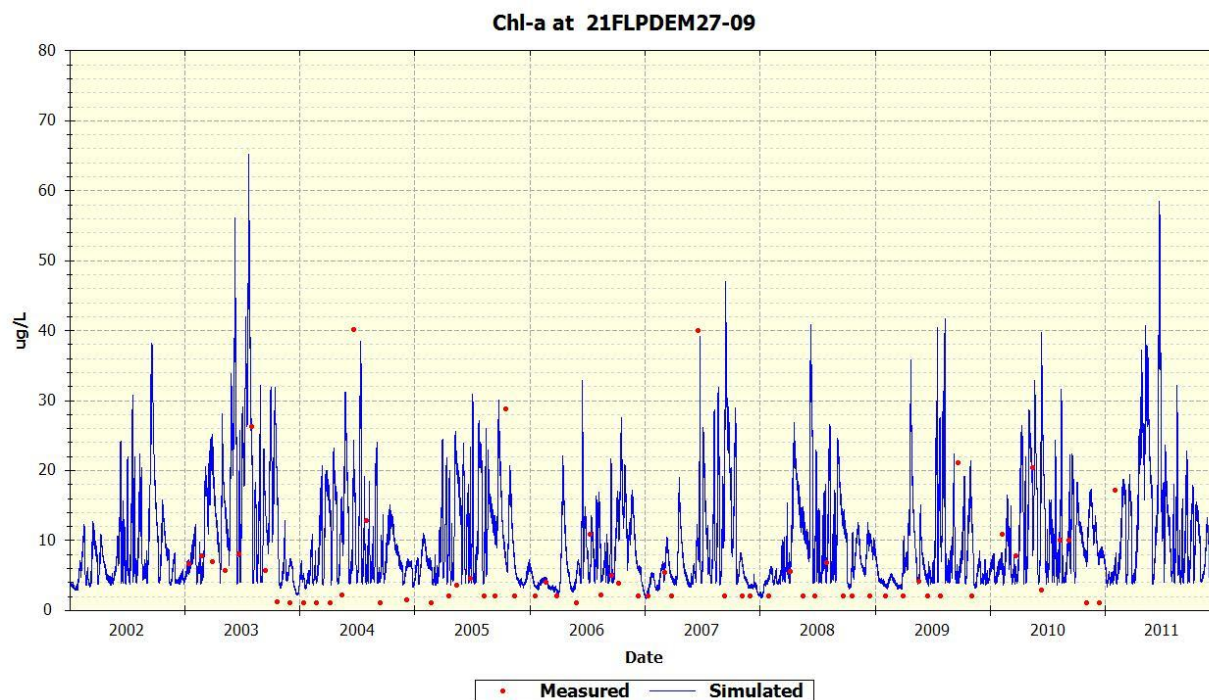


Figure 7.41 Measured verse modeled chlorophyll-a (ug/L) in McKay Creek at station 21FLPDEM27-09

Table 7.1 Current condition concentrations in the impaired WBID in the McKay Creek basin

Parameter	WBID 1633B
Total nitrogen (mg/L)	0.9
Total phosphorus (mg/L)	0.1
cBOD (mg/L)	1.6
DO (mg/L)	7.1

Table 7.2 Current condition loadings in the impaired WBID in the McKay Creek basin

Parameter	WBID 1633B	
	WLA (kg/yr)	LA (kg/yr)
Total nitrogen (mg/L)	--	6,838
Total phosphorus (mg/L)	--	896
BOD (mg/L)	--	16,196

## 7.2.2 Natural Condition

The natural condition scenario was developed to estimate water quality conditions if there was no impact from anthropogenic sources. The point sources located in the model were removed for the natural condition analysis. Land uses that were associated with anthropogenic activities (urban, agriculture, transportation, barren lands and rangeland) were converted to upland forests or forested wetlands based on the current ration of forest and wetland land uses in the model. Additionally, following the initial natural condition scenario run, sediment oxygen demand (SOD) was revised by using the following formula:  $SOD_{\text{revised}} = (\text{Avg Chla}_{\text{natural}} / \text{Avg Chla}_{\text{existing}}) * SOD$ . The lower, revised SOD represents the change expected in SOD following excessive nutrient removal from the system. The natural condition water quality predictions are presented in Table 7.3 and Table 7.4.

The purpose of the natural conditions scenario was to determine whether water quality standards could be achieved without abating the naturally occurring loads from the watershed. The natural condition modeling scenario indicated that the DO standard is achievable under natural conditions, indicating that low DO is not a naturally occurring phenomenon in WBID 1633B. Figure 7.43 through Figure 7.45 provide the natural condition scenario modeled parameters for WBID 1633B, and Figure 7.46 provides the cumulative distribution function of DO concentrations for both the modeled existing condition and natural condition results. The cumulative distribution curve shows there is an increase in DO concentrations in the natural condition scenario, specifically in DO concentration values less than 5 mg/L in the existing condition run.

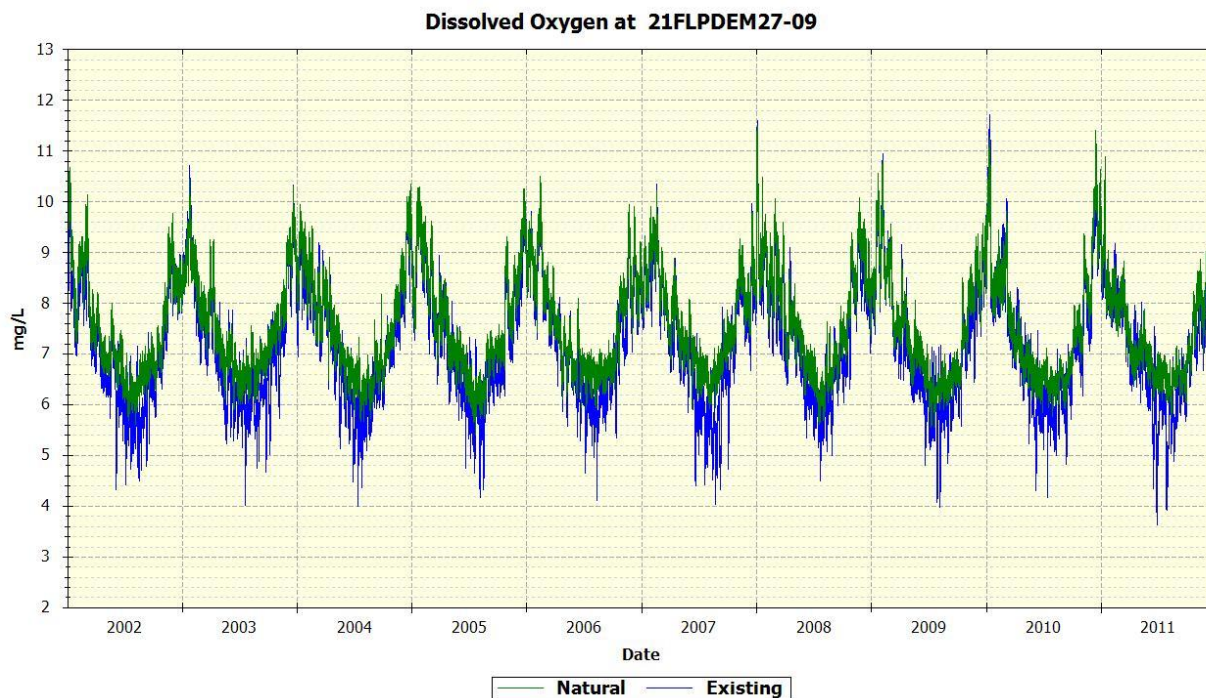


Figure 7.42 Existing condition versus natural condition dissolved oxygen (mg/L) in McKay Creek at station 21FLPDEM27-09

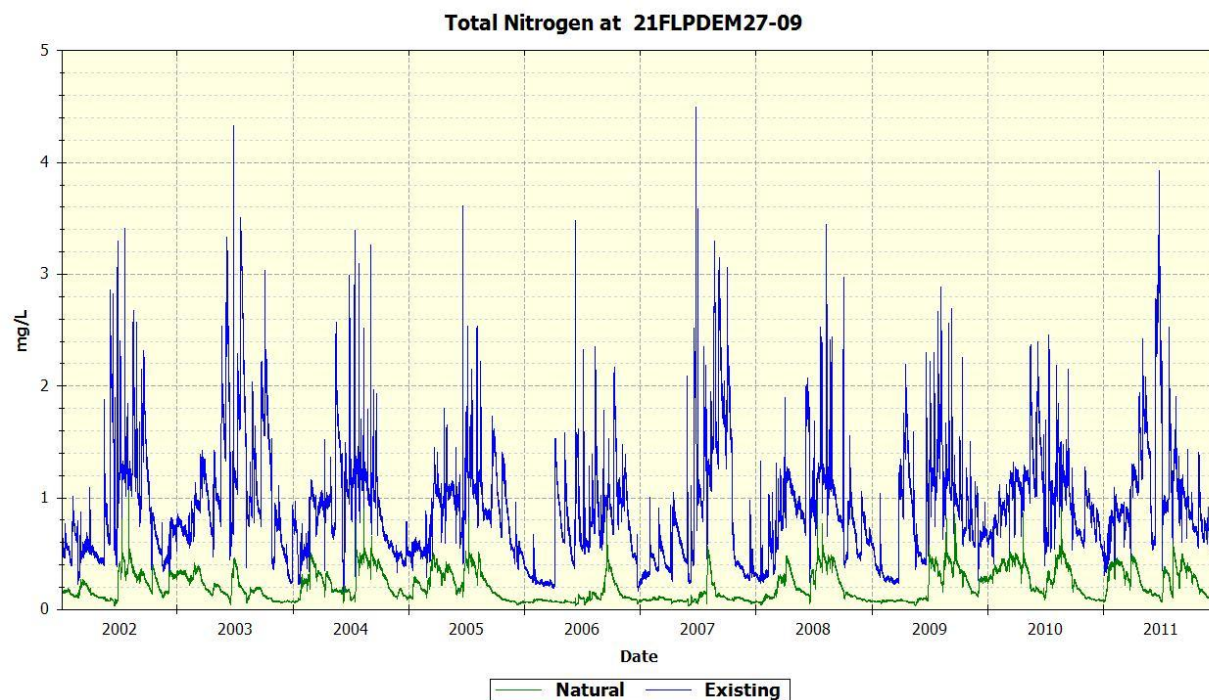


Figure 7.43 Existing condition versus natural condition total nitrogen (mg/L) in McKay Creek at station 21FLPDEM27-09

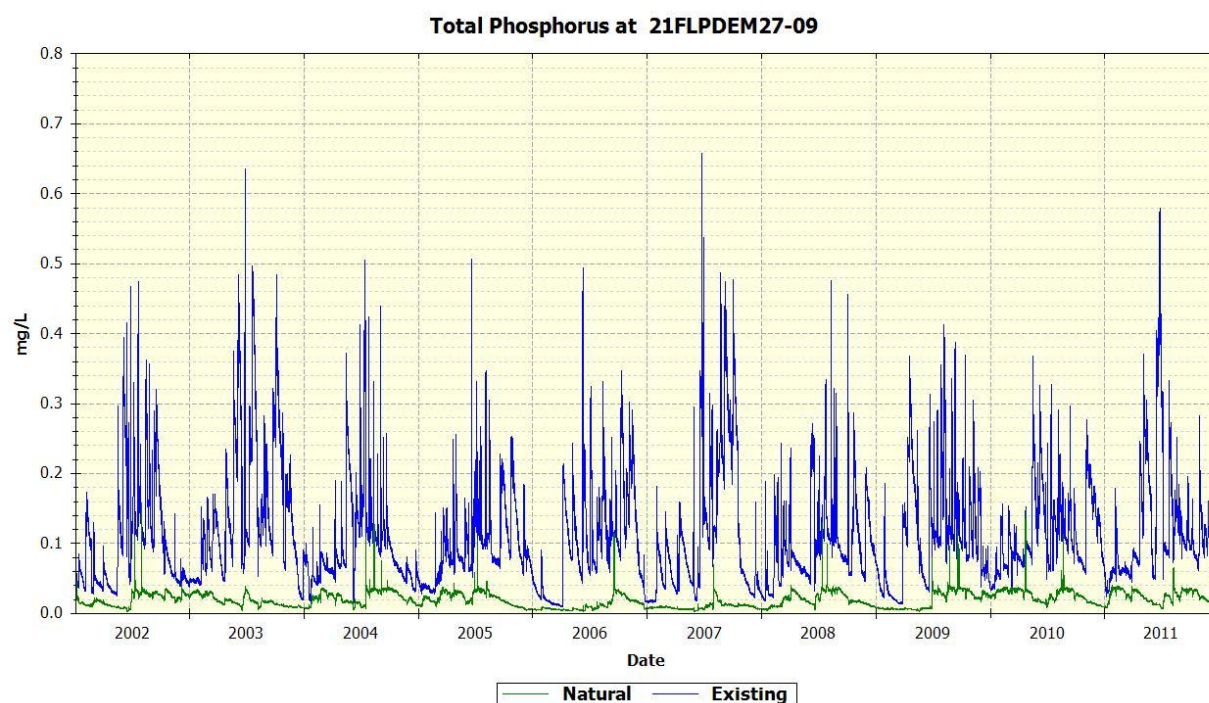


Figure 7.44 Existing condition versus natural condition total phosphorus (mg/L) in McKay Creek at station 21FLPDEM27-09

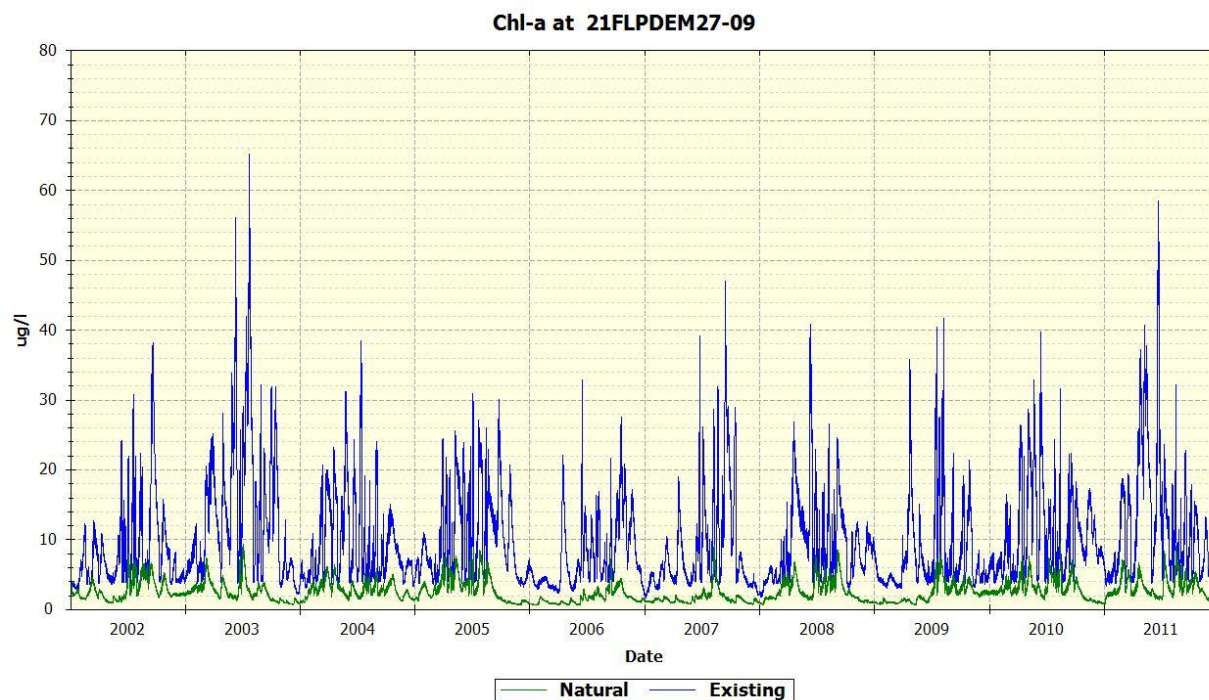


Figure 7.45 Existing condition verses natural condition chlorophyll a (ug/L) in McKay Creek at station 21FLPDEM27-09

Table 7.3 Natural condition concentrations in the impaired WBID in the McKay Creek basin

Parameter	WBID 1633B
Total nitrogen (mg/L)	0.2
Total phosphorus (mg/L)	0.0
cBOD (mg/L)	0.7
DO (mg/L)	7.5

Table 7.4 Natural condition loadings in the impaired WBID in the McKay Creek basin

Parameter	WBID 1633B	
	WLA (kg/yr)	LA (kg/yr)
Total nitrogen (mg/L)	--	1,473
Total phosphorus (mg/L)	--	135
BOD (mg/L)	--	5,806

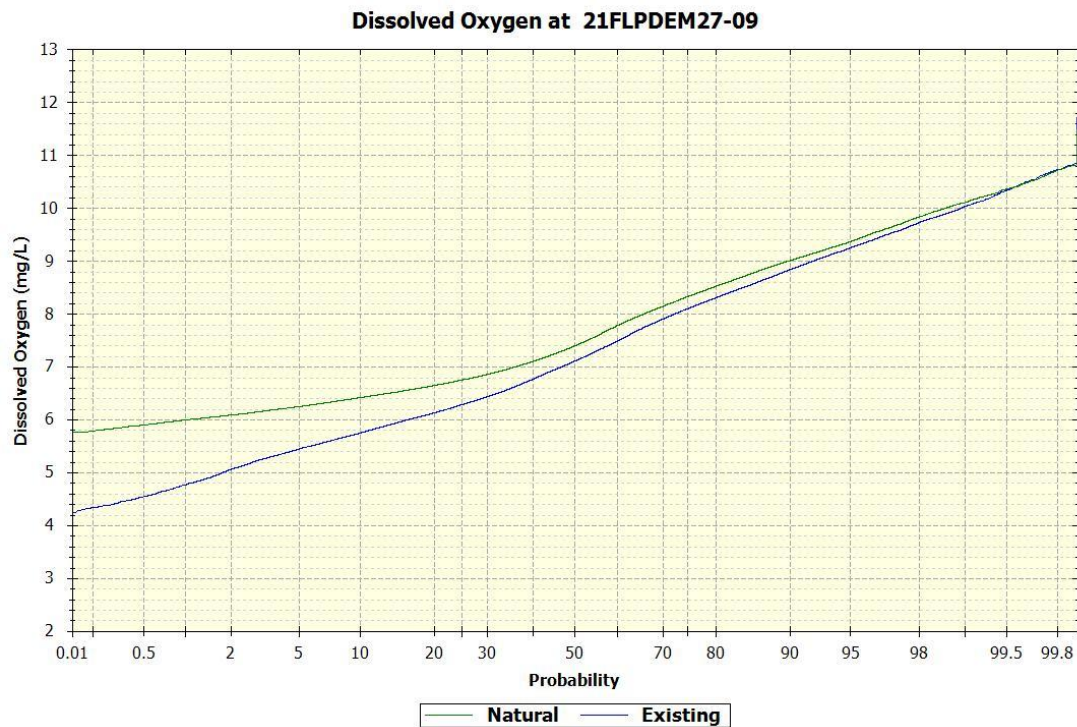


Figure 7.46 Dissolved oxygen concentration cumulative distribution function in McKay Creek at station 21FLPDEM27-09

### 7.2.3 Reduction Scenario

The natural condition scenario indicated that DO standards can be achieved in McKay Creek. A scenario was developed to determine what reductions in nutrients allowed for DO to be greater than 5 mg/L throughout the modeling period. A reduction in nutrients and biochemical oxygen demand of 35 percent from anthropogenic land uses raised DO concentrations above 5 mg/L at all times. The lowest modeled DO concentration at this reduction level was 5.02 mg/L. Figure 7.47 through Figure 7.50 show the results of the scenario reduction, while Table 7.5 and Table 7.6 provide averages and loads of water quality parameters in the scenario condition.

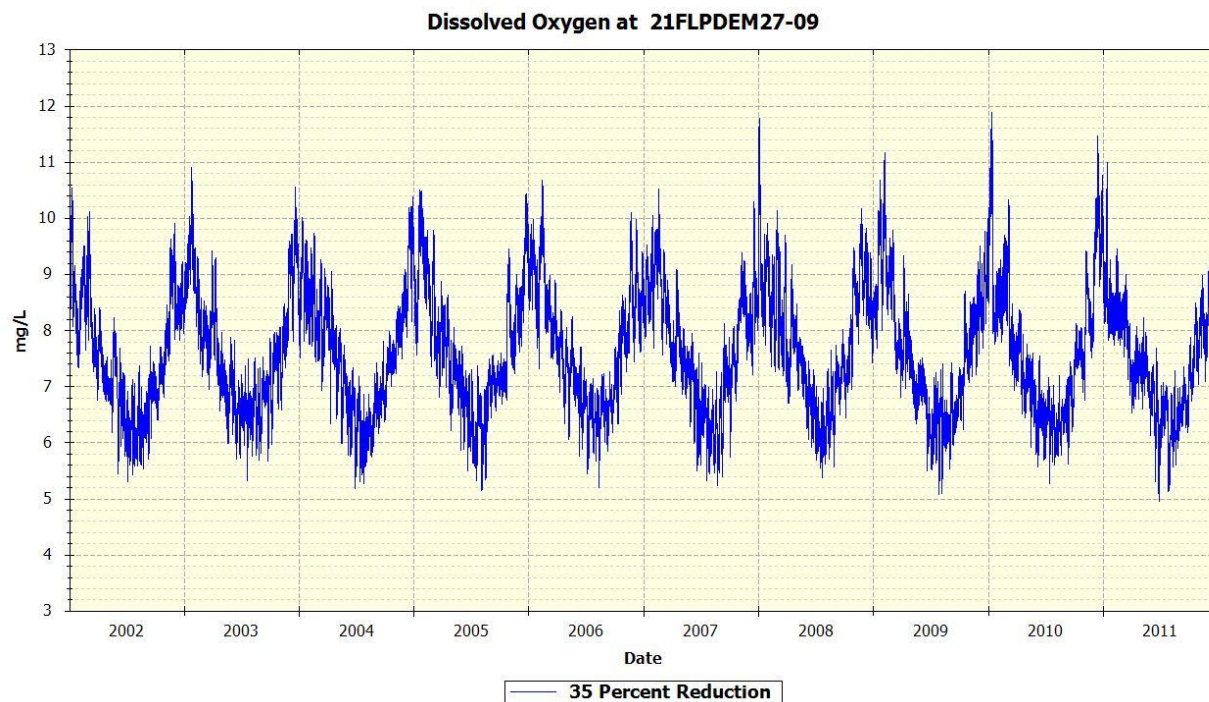


Figure 7.47 Reduction scenario for dissolved oxygen (mg/L) in McKay Creek at station 21FLPDEM27-09

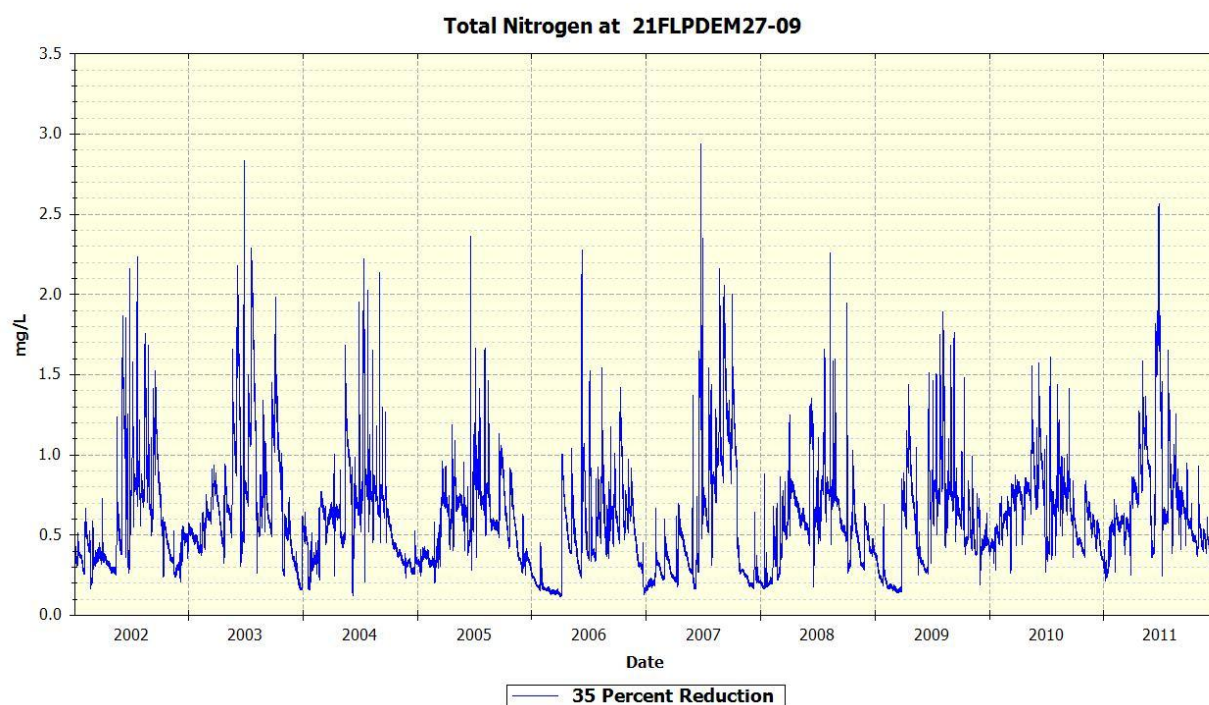


Figure 7.48 Reduction scenario for total nitrogen (mg/L) in McKay Creek at station 21FLPDEM27-09

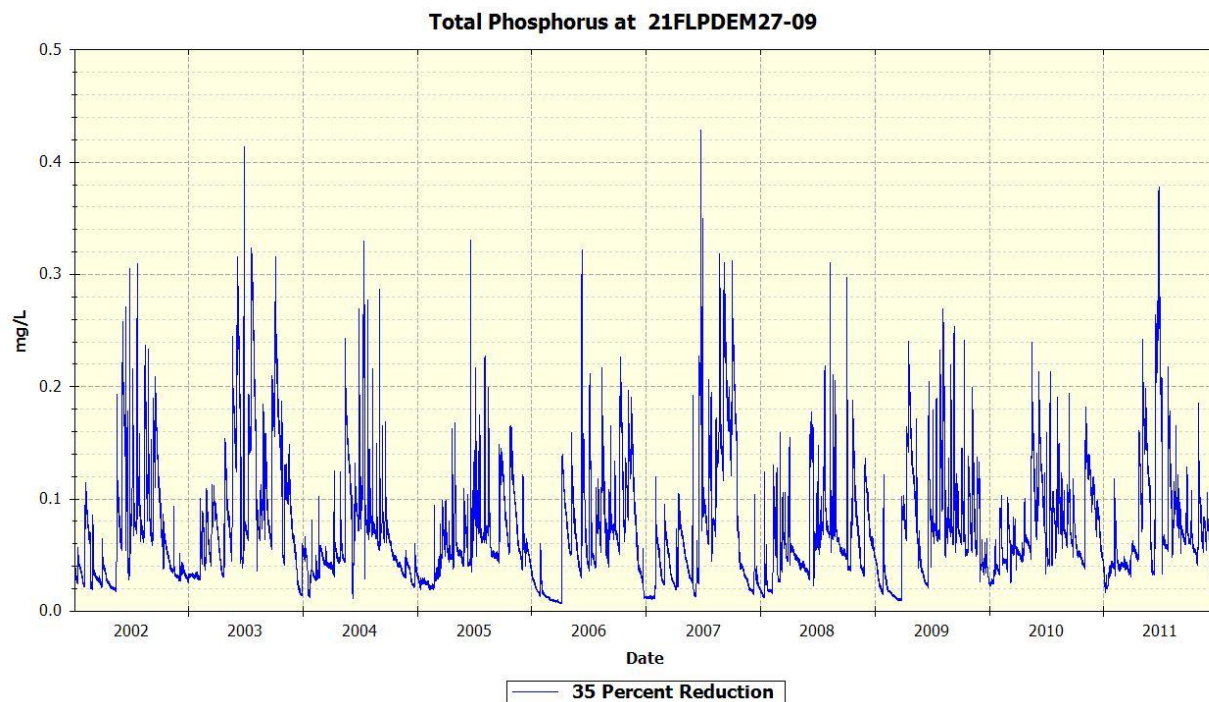


Figure 7.49 Reduction scenario for total phosphorus (mg/L) in McKay Creek at station 21FLPDEM27-09

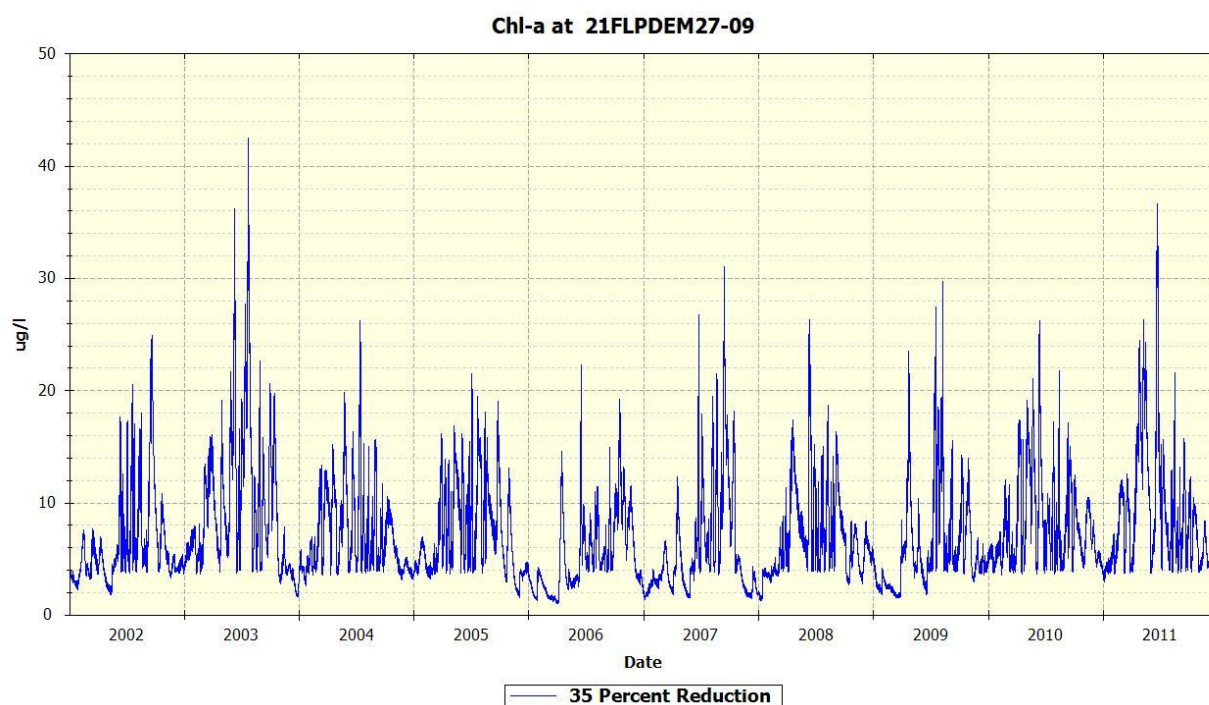


Figure 7.50 Reduction scenario for chlorophyll a (ug/L) in McKay Creek at station 21FLPDEM27-09

Table 7.5 Mean concentrations in the impaired WBID in the McKay Creek basin under the reduction scenario

Parameter	WBID 1633B
Total nitrogen (mg/L)	0.60
Total phosphorus (mg/L)	0.07
cBOD (mg/L)	1.09
DO (mg/L)	7.61

Table 7.6 Loadings in the impaired WBID in the McKay Creek basin under the reduction scenario

Parameter	WBID 1633B	
	WLA (kg/yr)	LA (kg/yr)
Total nitrogen (mg/L)	--	4,445
Total phosphorus (mg/L)	--	582
BOD (mg/L)	--	10,527

## 8.0 TMDL DETERMINATION

The TMDL for a given pollutant and waterbody is comprised of the sum of individual wasteload allocations (WLAs) for point sources, and load allocations (LAs) for both nonpoint sources and natural background levels. In addition, the TMDL must include a margin of safety (MOS), either implicitly or explicitly, to account for the uncertainty in the relationship between pollutant loads and the quality of the receiving waterbody. Conceptually, this definition is represented by the equation:

$$\text{TMDL} = \sum \square \text{WLAs} + \sum \square \text{LAs} + \text{MOS}$$

The TMDL is the total amount of pollutant that can be assimilated by the receiving waterbody and still achieve water quality standards and the waterbody's designated use. In this TMDL development, allowable concentrations from all pollutant sources that cumulatively amount to no more than the TMDL must be set and thereby provide the basis to establish water quality-based controls. These TMDLs are expressed as annual geometric mean concentrations, since the approach used to determine the TMDL targets relied on geometric means. The TMDLs targets were determined to be the conditions needed to restore and maintain a balanced aquatic system. Furthermore, it is important to consider nutrient loading over time, since nutrients can accumulate in waterbodies.

The TMDL was determined for the concentrations and loadings at the outlet of WBID 1633B, and included all loadings from upstream sources and streams. During the development of this TMDL, it was determined that the natural condition scenario (removal of all anthropogenic sources and land uses) did meet the Florida standards for DO A scenario was developed to determine what reductions in nutrients allowed for DO to be greater than 5 mg/L throughout the modeling period. A reduction in nutrients and biochemical oxygen demand of 35 percent from anthropogenic land uses raised DO concentrations above 5 mg/L at all times. The allocations for WBID 1633B for total nitrogen, total phosphorus, and biochemical oxygen demand are presented in Table 8.1.

Table 8.1 TMDL Load Allocations for McKay Creek, WBID 1633B

Constituent	Current Condition		TMDL Condition		Percent reduction		
	WLA (kg/yr)	LA (kg/yr)	WLA (kg/yr)	LA (kg/yr)	WLA	LA	MS4
Total Nitrogen	--	6,838	--	4,445	--	35%	35%
Total Phosphorus	--	896	--	582	--	35%	35%

### 8.1 Critical Conditions and Seasonal Variation

EPA regulations at 40 CFR 130.7(c)(1) require TMDLs to take into account critical conditions for stream flow, loading, and water quality parameters. The critical condition is the combination of environmental factors creating the "worst case" scenario of water quality conditions in the waterbody. By achieving the water quality standards at critical conditions, it is expected that water quality standards should be achieved during all other times. Seasonal variation must also be considered to ensure that water quality standards will be met during all seasons of the year, and that the TMDLs account for any seasonal change in flow or pollutant discharges, and any applicable water quality criteria or designated uses (such as swimming) that are expressed on a seasonal basis.

The critical condition for nonpoint source concentration and wet weather point source concentrations is typically an extended dry period followed by a rainfall runoff event. During the dry weather period, nutrients build up on the land surface, and are washed off by rainfall. The critical condition for continuous point source concentrations typically occurs during periods of low stream flow when dilution is minimized. Although loading of nonpoint source pollutants contributing to a nutrient impairment may occur during a runoff event, the expression of that nutrient impairment is more likely to occur during warmer months, and at times when the waterbody is poorly flushed.

## 8.2 Margin of Safety

The Margin of Safety accounts for uncertainty in the relationship between a pollutant load and the resultant condition of the waterbody. There are two methods for incorporating an MOS into TMDLs (USEPA 1991):

- Implicitly incorporate the MOS using conservative model assumptions to develop allocations
- Explicitly specify a portion of the total TMDL as the MOS and use the remainder for Allocations

This TMDL uses an implicit MOS since the TMDL targets for nutrients were set to natural background conditions.

## 8.3 Waste Load Allocations

Only MS4s and NPDES facilities discharging directly into lake segments (or upstream tributaries of those segments) are assigned a WLA. The WLAs, if applicable, are expressed separately for continuous discharge facilities (e.g., WWTPs) and MS4 areas, as the former discharges during all weather conditions whereas the later discharges in response to storm events.

### 8.3.1 Wastewater/Industrial Permitted Facilities

A TMDL wasteload allocation (WLA) is given to wastewater and industrial NPDES-permitted facilities discharging to surface waters within an impaired watershed. There are no continuous discharge NPDES-permitted point sources in WBID 1633B, therefore no WLA was calculated.

### 8.3.2 Municipal Separate Storm Sewer System Permits

The WLA for MS4 FLS000005 and FLS000007 are expressed in terms of percent reductions equivalent to the reductions required for nonpoint sources. Given the available data, it is not possible to estimate concentrations coming exclusively from the MS4 areas. Although the aggregate concentration allocations for stormwater discharges are expressed in numeric form, i.e., percent reduction, based on the information available today, it is infeasible to calculate numeric WLAs for individual stormwater outfalls because discharges from these sources can be highly intermittent, are usually characterized by very high flows occurring over relatively short time intervals, and carry a variety of pollutants whose nature and extent varies according to geography and local land use. For example, municipal sources such as those covered by this TMDL often include numerous individual outfalls spread over large areas. Water quality impacts, in turn, also depend on a wide range of factors, including the magnitude and duration of rainfall events, the time period between events, soil conditions, fraction of land that is impervious to rainfall, other land use activities, and the ratio of stormwater discharge to receiving water flow.

This TMDL assumes for the reasons stated above that it is infeasible to calculate numeric water quality-based effluent limitations for stormwater discharges. Therefore, in the absence of information presented to the permitting authority showing otherwise, this TMDL assumes that water quality-based effluent limitations for stormwater sources of nutrients derived from this

TMDL can be expressed in narrative form (e.g., as best management practices), provided that: (1) the permitting authority explains in the permit fact sheet the reasons it expects the chosen BMPs to achieve the aggregate wasteload allocation for these stormwater discharges; and (2) the state will perform ambient water quality monitoring for nutrients for the purpose of determining whether the BMPs in fact are achieving such aggregate wasteload allocation.

All Phase 1 MS4 permits issued in Florida include a re-opener clause allowing permit revisions for implementing TMDLs once they are formally adopted by rule. Florida may designate an area as a regulated Phase II MS4 in accordance with Rule 62-620.800, FAC. Florida's Phase II MS4 Generic Permit has a "self-implementing" provision that requires MS4 permittees to update their stormwater management program as needed to meet their TMDL allocations once those TMDLs are adopted. Permitted MS4s will be responsible for reducing only the loads associated with stormwater outfalls which it owns, manages, or otherwise has responsible control. MS4s are not responsible for reducing other nonpoint source loads within its jurisdiction. All future MS4s permitted in the area are automatically prescribed a WLA equivalent to the percent reduction assigned to the LA. The MS4 service areas described in Section 6.1.2 of this report are required to meet the percent reduction prescribed in Table 8.1 through the implementation of BMPs.

#### **8.4 Load Allocations**

The load allocation for nonpoint sources was assigned a percent reduction in nutrient concentrations from the current concentrations coming into the WBID addressed in the TMDL report.

### **9.0 RECOMMENDATIONS/IMPLEMENTATION**

The initial step in implementing a TMDL is to more specifically locate pollutant source(s) in the watershed. FDEP employs the Basin Management Action Plan (B-MAP) as the mechanism for developing strategies to accomplish the specified load reductions. Components of a B-MAP are:

- Allocations among stakeholders
- Listing of specific activities to achieve reductions
- Project initiation and completion timeliness
- Identification of funding opportunities
- Agreements
- Local ordinances
- Local water quality standards and permits
- Follow-up monitoring

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